

Canadian Biodiversity: Ecosystem Status and Trends 2010
Evidence for Key Findings Summary Report No. 3
Published by the Canadian Councils of Resource Ministers





Library and Archives Canada Cataloguing in Publication

Atlantic Maritime Ecozone<sup>+</sup> evidence for key findings summary.

Issued also in French under title:

Sommaire des éléments probants relativement aux constatations clés pour l'écozone<sup>+</sup> maritime de l'Atlantique.

Electronic monograph in PDF format.

En14-43/0-3-2015E-PDF

978-1-100-23826-5

Information contained in this publication or product may be reproduced, in part or in whole, and by any means, for personal or public non-commercial purposes, without charge or further permission, unless otherwise specified.

You are asked to:

- Exercise due diligence in ensuring the accuracy of the materials reproduced;
- Indicate both the complete title of the materials reproduced, as well as the author organization; and
- Indicate that the reproduction is a copy of an official work that is published by the Government of Canada and that the reproduction has not been produced in affiliation with or with the endorsement of the Government of Canada.

Commercial reproduction and distribution is prohibited except with written permission from the author. For more information, please contact Environment Canada's Inquiry Centre at 1-800-668-6767 (in Canada only) or 819-997-2800 or email to enviroinfo@ec.gc.ca.

Cover photos: Margaree Valley, Cape Breton Island, Nova Scotia, © iStock.com / cworthy; Hopewell Rocks, Bay of Fundy, New Brunswick, © iStock.com / MorganLeFaye

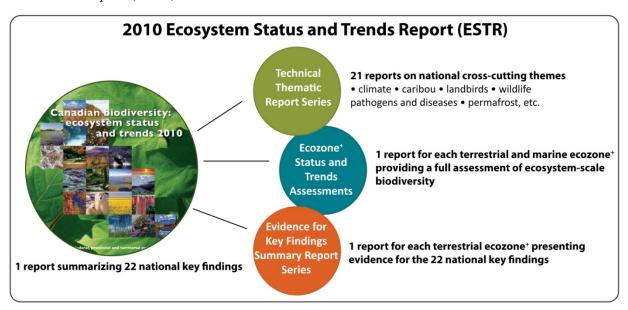
This report should be cited as:

ESTR Secretariat. 2014. Atlantic Maritime Ecozone<sup>+</sup> evidence for key findings summary. Canadian Biodiversity: Ecosystem Status and Trends 2010, Evidence for Key Findings Summary Report No. 3. Canadian Councils of Resource Ministers. Ottawa, ON. ix + 100 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-1">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-1</a>

© Her Majesty the Queen in Right of Canada, 2014 Aussi disponible en français

## **PREFACE**

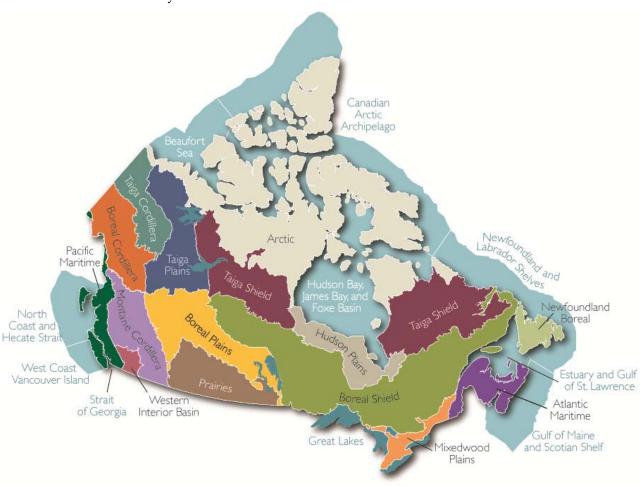
The Canadian Councils of Resource Ministers developed a Biodiversity Outcomes Framework¹ in 2006 to focus conservation and restoration actions under the *Canadian Biodiversity Strategy*.² *Canadian Biodiversity: Ecosystem Status and Trends 2010*³ was the first report under this framework. It presents 22 key findings that emerged from synthesis and analysis of background technical reports prepared on the status and trends for many cross-cutting national themes (the Technical Thematic Report Series) and for individual terrestrial and marine ecozones⁺ of Canada (the Ecozone⁺ Status and Trends Assessment Report Series). More than 500 experts participated in data analysis, writing, and review of these foundation documents. Summary reports were also prepared for each terrestrial ecozone⁺ to present the ecozone⁺ specific evidence related to each of the 22 national key findings (the Evidence for Key Findings Summary Report Series). Together, the full complement of these products constitutes the 2010 Ecosystem Status and Trends Report (ESTR).



This report, *Atlantic Maritime Ecozone*<sup>+</sup> *Evidence for Key Findings Summary*, presents evidence from the *Atlantic Maritime Ecozone*<sup>+</sup> *Status and Trends Assessment*<sup>4</sup> related to the 22 national key findings and highlights important trends specific to this ecozone<sup>+</sup>. It is not a comprehensive assessment of all ecosystem-related information. The level of detail presented on each key finding varies and important issues or datasets may have been missed. Also, because of the report's timing or a lack of readily available ecozone<sup>+</sup>-specific information, some key findings were not addressed. Some emphasis has been placed on information from the national Technical Thematic Report Series. As in all ESTR products, the time frames over which trends are assessed vary—both because time frames that are meaningful for these diverse aspects of ecosystems vary and because the assessment is based on the best available information, which is over a range of time periods.

# Ecological classification system – ecozones<sup>†</sup>

A slightly modified version of the Terrestrial Ecozones of Canada, described in the *National Ecological Framework for Canada*,<sup>5</sup> provided the ecosystem-based units for all reports related to this project. Modifications from the original framework include: adjustments to terrestrial boundaries to reflect improvements from ground-truthing exercises; the combination of three Arctic ecozones into one; the use of two ecoprovinces—Western Interior Basin and Newfoundland Boreal; the addition of nine marine ecosystem-based units; and, the addition of the Great Lakes as a unit. This modified classification system is referred to as "ecozones" throughout these reports to avoid confusion with the more familiar "ecozones" of the original framework.<sup>6</sup> The boundary for the Atlantic Maritime is the same in both frameworks.



# **Acknowledgements**

The ESTR Secretariat acknowledges Trish Hayes, Dan Beaudette (New Brunswick Department of Natural Resources), and Greg Sheehy for the preparation of various drafts of this report. The report was overseen and edited by Trish Hayes and Patrick Lilley. Kelly Badger was the lead graphics designer. Additional support was provided by Jodi Frisk, Isabelle Turcotte, Eric Jacobsen, Ellorie McKnight, Michelle Connolly, and others. It is based on the *Atlantic Maritime Ecozone*<sup>+</sup> *Status and Trends Assessment*.<sup>4</sup> Other experts made significant contributions to that draft report and are listed below. Reviews were provided by scientists and resource managers from relevant provincial/territorial and federal government agencies. The Canadian Society of Ecology and Evolution also coordinated reviews with external experts.

## Atlantic Maritime Ecozone<sup>†</sup> Draft Status and Trends Assessment acknowledgements

Lead author: S. Eaton

Contributing authors: J. Barr, D. Beaudette, T. Hayes

Contributing authors, specific sections or topics: Ecosystem services: G. MacAskill

Climate change impacts in St. Lawrence: J.-P. Savard and R. Siron

#### Authors of ESTR Technical Thematic Reports from which material is drawn:

Canadian climate trends, 1950-2007: X. Zhang, R. Brown, L. Vincent, W. Skinner, Y. Feng and E. Mekis Trends in large fires in Canada, 1959-2007: C.C. Krezek-Hanes, F. Ahern, A. Cantin and M.D. Flannigan Wildlife pathogens and diseases in Canada: F.A. Leighton

Trends in breeding waterfowl in Canada: M. Fast, B. Collins and M. Gendron Landbird trends in Canada, 1968-2006: C. Downes, P. Blancher and B. Collins

Trends in Canadian shorebirds: C. Gratto-Trevor, R.I.G. Morrison, B. Collins, J. Rausch and V. Johnston Trends in wildlife habitat capacity on agricultural land in Canada, 1986-2006: S.K. Javorek and M.C. Grant

Trends in residual soil nitrogen for agricultural land in Canada, 1981-2006: C.F. Drury, J.Y. Yang and R. De Jong

Soil erosion on cropland: introduction and trends for Canada: B.G. McConkey, D.A. Lobb, S. Li, J.M.W. Black and P.M. Krug

Monitoring biodiversity remotely: a selection of trends measured from satellite observations of Canada: F. Ahern, J. Frisk, R. Latifovic and D. Pouliot

Inland colonial waterbird and marsh bird trends for Canada: D.V.C. Weseloh

Climate-driven trends in Canadian streamflow, 1961-2003: A. Cannon, T. Lai and P. Whitfield Biodiversity in Canadian lakes and rivers: W.A. Monk and D.J. Baird

**Review** conducted by scientists and renewable resource and wildlife managers from provincial and federal government agencies through a review process administered by the ESTR Steering Committee. Additional reviews of specific sections were conducted by external experts in their field of expertise.

**Direction** provided by the ESTR Steering Committee composed of representatives of federal, provincial, and territorial agencies.

**Editing, synthesis, technical contributions, maps and graphics, and report production** by the ESTR Secretariat of Environment Canada.

Aboriginal Traditional Knowledge compiled from publicly available sources by D.D. Hurlburt.

# **Table of Contents**

PREFACE	
Ecological classification system – ecozones <sup>†</sup>	i
Acknowledgements	ii
ECOZONE <sup>+</sup> BASICS	2
KEY FINDINGS AT A GLANCE: NATIONAL AND ECOZONE <sup>+</sup> LEVEL	6
THEME: BIOMES	13
Forests	13
Forest age structure	15
Acadian forest	16
Forest subdomains in Quebec portion of AME	17
Forest composition	18
Acadian forest	18
Forest subdomains in Quebec portion of AME	19
Fragmentation	21
Forest birds	21
Cumulative human impact	22
Wetlands	
Waterfowl	24
Lakes and rivers	25
Streamflow in natural rivers	25
Water control structures	27
Coastal	29
Coastal wetlands	29
Eelgrass	30
Shorebirds	
Waterfowl	33
Sandy shores and sand dunes	33
Piping plover	34
Coastal development	
Sea-level rise and coastal erosion	
Ice across biomes	37
River and lake ice	37
Sea ice	38
THEME: HUMAN/ECOSYSTEM INTERACTIONS	40
Protected areas	
Invasive non-native species	42
Invasive plants	42
Invasive non-native insects and diseases	
Beech bark disease	44
Brown spruce longhorn beetle	
Invasive non-native freshwater species	
Smallmouth bass	
Didymo	45
Invasive non-native marine species	

European green crab	46
Nutrient loading and algal blooms	46
Nitrogen	46
Nitrate levels in groundwater and surface water in PEI	47
Phosphorous	50
Phosphorus concentrations in rivers in Quebec	50
Algal blooms in Quebec	51
Acid deposition	52
Climate change	
Trends in climatic variables	54
Future climate predictions	59
Ecosystem services	
THEME: HABITAT, WILDLIFE, AND ECOSYSTEM PROCESSES	60
Agricultural landscapes as habitat	
Wildlife habitat capacity on agricultural land	62
Soil erosion on cropland	
Birds of grassland and other open habitats	
Species of special economic, cultural, or ecological interest	
Woodland caribou	66
Other ungulates	66
Atlantic salmon	67
American eel	69
Freshwater fish	69
Landbirds	70
Primary productivity	71
Natural disturbance	73
Extreme weather events	73
Tropical storms and hurricanes	73
Storm surges and flooding	
Fire	
Large-scale native insect outbreaks	76
Spruce budworm	
Food webs	
THEME: SCIENCE/POLICY INTERFACE	80
Biodiversity monitoring, research, information management, and reporting	
Rapid change and thresholds	
CONCLUSION: HUMAN WELL-BEING AND BIODIVERSITY	82
REFERENCES	83

# List of Figures

Figure 1. Overview map of the Atlantic Maritime Ecozone <sup>+</sup>	1
Figure 2. Land cover of the Atlantic Maritime Ecozone⁺, 2005	3
Figure 3. Human population of the Atlantic Maritime Ecozone <sup>+</sup> , 1971–2006	4
Figure 4. Forest regions and the principal tree species within each region	14
Figure 5. Quebec forest domains	15
Figure 6. Age-class distribution by forest type on stocked forest land as a percentage of the total stocked forest area in the Atlantic Maritime Ecozone⁺, 1999	16
Figure 7. Percent of total forest area in each age class in Nova Scotia, 1958–2003	17
Figure 8. Proportion of forest at each major developmental stage in Quebec's subdomains that occur in the Atlantic Maritime Ecozone <sup>+</sup> , 1970s, 1980s, and 1990s 1990s	18
Figure 9. Estimated frequency of major forest tree genera in Kings County, NB, 1800 and 1993	19
Figure 10. Proportion of total area covered by different forest cover types in the Quebec subdomains that occur in the Atlantic Maritime Ecozone <sup>+</sup> , 1970s, 1980s, and 1990s	20
Figure 11. Change in forest types for the Appalachian Ecoregion in southern Quebec between 1993 and 2001	20
Figure 12. Intact landscape fragments larger than 50 km² in the Atlantic Maritime Ecozone <sup>+</sup> , 2003	21
Figure 13. Annual indices of population change for birds of forest habitat (left) and shrub-early successional habitat (right), 1968–2006	22
Figure 14. The human footprint of the Northern Appalachian/Acadian Ecoregion, 2008	23
Figure 15. Summary of the total number of sites displaying increasing and decreasing trends in various streamflow variables in the Atlantic Maritime Ecozone⁺, 1970–2005	26
Figure 16. Changes in streamflow between 1961–1982 and 1983–2003 for representative sites of each hydrology group in the Atlantic Maritime Ecozone <sup>+</sup>	27
Figure 17. Spatial distribution of dams greater than 10 m in height within the Atlantic Maritime Ecozone <sup>†</sup> , grouped by year of completion between 1830 and 2005	28
Figure 18. Decline in area of vegetated salt marsh in five locations in southeastern New Brunswick between 1944 and 2001	30
Figure 19. Decline in area of beach and dune habitat in five locations in southeastern New  Brunswick between 1944 and 2001	34
Figure 20. Distribution of 2006 piping plover nesting sites (left, map) and the number of piping plover adults counted during surveys (right, bar chart) in the Atlantic Maritime Ecozone <sup>+</sup> , 1991, 1996, 2001, and 2006.	35
Figure 21. Trends in lot registration within two km of the Nova Scotia coastline by decade	35
Figure 22. Trend in annual mean water level in six harbours in the Atlantic Maritime Ecozone⁺	36
Figure 23. Trend in total accumulated ice coverage (top) and length of season (cover >10%)	
(bottom) for the Gulf of St. Lawrence, 1971–2005	
Figure 24. Distribution of protected areas in the Atlantic Maritime Ecozone <sup>+</sup> , May 2009	
Figure 25. Growth of protected areas in the Atlantic Maritime Ecozone <sup>+</sup> , 1936–2009	41

Figure 26	. Total number of native and non-native plant species in the Maritime provinces, 20014	2
Figure 27	. Abundance of non-native plant species in the Maritime Provinces, 20014	.3
Figure 28	. Number of lakes with first known occurrences of smallmouth bass in Nova Scotia,	_
	1942–2008	.5
Figure 29	. Change in Residual Soil Nitrogen (RSN) risk class from 1981 to 2006 (left) and risk classes in 2006 (right) for agricultural land in the Atlantic Maritime Ecozone <sup>†</sup> 4	7
•	. Mean nitrate levels and the percentage of private wells exceeding recommended nitrate concentrations, PEI, 1984–2007. There were no data from 1985–1994 and 1996–1999	-8
Figure 31	. Change in average groundwater nitrate concentrations between 2000–2005 (top) and 2004–2008 (bottom) in watersheds in PEI4	.9
Figure 32	. Number of anoxic events reported on Prince Edward Island between 2002 and 2008 5	0
Figure 33	. Risk of water contamination by phosphorous in agricultural watersheds under 2006 management practices in the Atlantic Maritime Ecozone <sup>+</sup> (map) and trend in the proportion of farmland in each risk class, 1981–2006, by province (bar graphs)5	1
Figure 34	. Number of lakes and rivers where blue-green algae was detected for Quebec administrative units that overlap with the Atlantic Maritime Ecozone <sup>+</sup> , 2004–20085	2
Figure 35	. Trend in wet sulphate and wet nitrate deposition in the Atlantic Maritime Ecozone $^+$ , 1990–1994 and 2000–20045	3
_	. Map of forest areas in the New England states and eastern Canadian provinces where critical load has been exceeded due to acid deposition, ca. 1999–2003	3
	. Change in mean temperatures in the Atlantic Maritime Ecozone <sup>+</sup> , 1950–2007, for: a) spring (March–May), b) summer (June–August, c) fall (September–November), and d) winter (December–February)	6
_	. Change in the amounts of precipitation in the Atlantic Maritime Ecozone <sup>+</sup> , 1950–2007, for: a) spring (March–May), b) summer (June–August, c) fall (September–November), and d) winter (December–February)5	7
Ū	. Change in snow durations (the number of days with ≥2 cm of snow on the ground) in the Atlantic Maritime Ecozone <sup>+</sup> , 1950–2007, in: a) the first half of the snow season (August–January), which indicates change in the start date of snow cover, and b) the second half of the snow season (February–July), which indicates changes in the end date of snow cover	8
	. Percentage of land defined as agricultural in the Atlantic Maritime Ecozone $^{\scriptscriptstyle +}$ , 2006 $6$	
Figure 41	. Total agriculture area, the amount of land per cover type (bar chart), and the relative percentage of each cover type (table) for the Atlantic Maritime Ecozone <sup>†</sup> in 1986, 1996, and 2006	
	The share of agricultural land in each habitat capacity category (left axis, stacked bars) and the average habitat capacity (right axis, points and line) for the Atlantic Maritime Ecozone <sup>+</sup> in 1986, 1996, and 2006.	
Figure 43	. Change in wildlife habitat capacity on agricultural lands in the Atlantic Maritime Ecozone <sup>+</sup> , 1986–20066	
	Soil erosion risk classes for cropland in the Atlantic Maritime Ecozone <sup>†</sup> , 20066	

rigure 43	5. Annual indices of population change in bird assemblages for grassland habitat (left) and other open habitats (right) in the Atlantic Maritime Ecozone <sup>+</sup> , 1968–2006	65
Figure 46	5. Trend in the estimated numbers of the Gaspésie woodland caribou population, 1983–2006	
Figure 47	7. Changes in abundance of salmon populations for the Maritime provinces (top) and Quebec (bottom), 1987–2005	68
Figure 48	3. Trend in numbers of imperilled freshwater and diadromous fish taxa in each status category for North American ecoregions in the Atlantic Maritime Ecozone <sup>+</sup> , 1979, 1989, and 2008.	70
Figure 49	9. Change in the Normalized Difference Vegetation Index for the Atlantic Maritime Ecozone <sup>+</sup> , 1985–2006	72
Figure 50	D. Trends in the average number of tropical cyclones in the Atlantic Basin, 1900–1999, 1950–1999, and 1991–2000	73
Figure 51	Storm surge maxima return level on the Atlantic coast of Canada based on the 40- year hindcast	74
Figure 52	2. Average number of storm surges per year above the threshold, by decade, at Charlottetown, PEI, 1940s—1990s	75
Figure 53	3. Total annual area burned by large fires (>2 km² in size) from 1959 to 2007 (left) and distribution of large fires from the 1980s to present in the Atlantic Maritime Ecozone <sup>+</sup>	76
Figure 54	1. Intervals of spruce budworm outbreaks in eastern Quebec identified during previous reconstructions based on tree-ring chronologies, 1500–2000	77
Figure 55	5. Trend in (A) spruce budworm defoliation in New Brunswick, 1949–2007, and (B) the area treated with pesticides, 1952–2007	78
Figure 56	5. Chronology of colonization of a portion of the Atlantic Maritime Ecozone <sup>+</sup> by the eastern coyote from the 1960s to the 1980s	79
List of Ta	bles	
Table 1.	Atlantic Maritime Ecozone <sup>†</sup> overview	2
Table 2.	Key findings overview	6
Table 3.	Abundance trends for selected breeding waterfowl species in the Atlantic Maritime Ecozone <sup>+</sup> , 1990s–2000s	24
Table 4.	Trends in eelgrass from studies in Nova Scotia and the Gulf of the St. Lawrence	31
Table 5.	Trends in abundance of shorebirds migrating through coastal areas of the Atlantic  Maritime Ecozone <sup>+</sup> , 1974–2006	32
Table 6.	Trends in abundance of coastal breeding shorebirds, 1970s to 2000s	33
Table 7.	Trends in lake freeze-up and break-up dates from studies in the Atlantic Maritime Ecozone <sup>+</sup>	38
Table 8.	Major invasive non-native insects, and diseases in Nova Scotia, including year of introduction, location of first introduction to North America, and preferred host species, 1890s–2000.	44

Table 9. Summary of changes in climate variables in the Atlantic Maritime Ecozone <sup>+</sup> , 1950− 2007	55
Table 10. Summary of the estimated values of ecosystem goods and services in the Atlantic Maritime Ecozone <sup>+</sup> , excluding the Quebec portion	60
Table 11. Identification and status of imperilled freshwater and diadromous fish taxa in the Atlantic Maritime Ecozone <sup>+</sup> , 1979, 1989, and 2008	70
Table 12. Trends in abundance of landbirds for the Atlantic Maritime Ecozone <sup>+</sup> , 1970s to 2000s	71

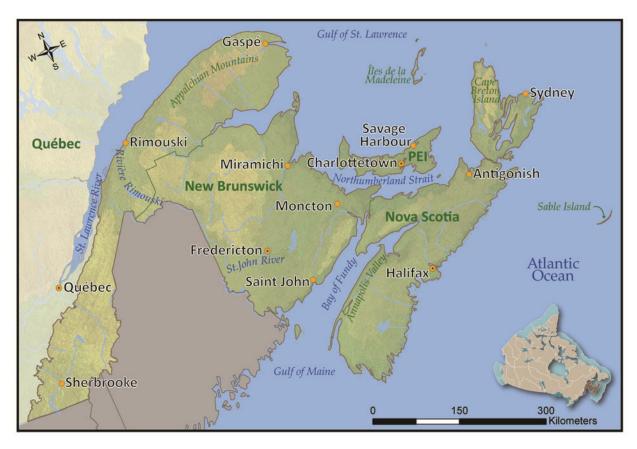


Figure 1. Overview map of the Atlantic Maritime  $Ecozone^{+}$ .

# **ECOZONE**<sup>†</sup> BASICS

The Atlantic Maritime Ecozone<sup>+</sup> (AME) (Figure 1) is located on the southern Atlantic coastline of Canada and fully encompasses the three Canadian Maritime provinces as well as a portion of southern Quebec (see national map on page ii).

The AME is a diverse landscape characterized by several types of forest, rocky shorelines, agricultural lands, lakes, and rivers. The ocean's proximity has a tremendous influence on the area's physical features and climate and plays an important role in shaping its ecosystems. Table 1 provides a summary of the main features of the ecozone<sup>+</sup>.

Table 1. Atlantic Maritime Ecozone<sup>+</sup> overview.

Area	205,836 km² (2.1% of Canada)	
Topography	Dominated by two features: the northern extension of the Appalachian	
	Mountains and the coastal lowlands of the Northumberland Plain	
Climate	Cool, moist maritime climate	
River basins	St. Lawrence, St. Mary's, and Miramichi rivers flowing to the Atlantic Ocean Saint John River is the largest inland river	
Geology	Landscape built by millions of years of volcanic and tectonic activity, mountain building, erosion, sedimentation, and several major glaciations Mix of sedimentary and igneous bedrock Surficial materials are 70% till	
Settlement	Majority of population located along low-lying coast Major settlements include Halifax, Saint John, Moncton, Fredericton, Charlottetown, Rimouski, and Sherbrooke	
Economy	Resource-based industry (forestry, agriculture, fishing, mining) Service industry Some manufacturing	
Development	Intensive development limited to major coastal communities Oil and gas development increasing offshore	
National/global significance	Seven Canadian national parks or national park reserves (R): Cape Breton Highlands, Prince Edward Island, Fundy, Kejimkujik, Kouchibouguac, Forillon, and Sable Island (R)  Thirteen National Wildlife Areas: Boot Island, John Lusby Marsh, Chignecto, Sand Pond, Sea Wolf Island, Wallace Bay, Cape Jourimain, Portage Island, Portobello Creek, Shepody, Tintamarre, Pointe-au-Père, and Îles de l'Estuaire Two biosphere reserves: Fundy and Southwest Nova  Eight Ramsar sites (wetlands of international significance): Baie de l'Isle-Verte, Chignecto, Musquodoboit Harbour, Southern Bight Minas Basin, Malpeque Bay, Mary's Point, Shepody Bay, and Tabusintac Lagoon and River Estuary Three Western Hemisphere Shorebird Reserve Network sites: Mary's Point, Shepody Bay, and Southern Bight—Minas Basin  One World Heritage Site: Miguasha National Park of Quebec (Devonian fossil beds)  Highest tides in the world in the Bay of Fundy	

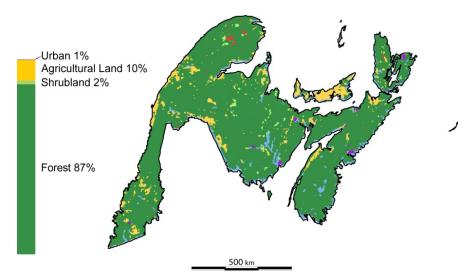


Figure 2. Land cover of the Atlantic Maritime Ecozone<sup>+</sup>, 2005. Source: Ahern et al., 2011 al., 2005<sup>8</sup>

**Jurisdictions:** The AME includes the provinces of New Brunswick (NB), Nova Scotia (NS), and Prince Edward Island (PEI), and the Gaspé Peninsula, Îles-de-la-Madeleine, and part of the southern shore of the St. Lawrence River in Quebec. Major Aboriginal groups in this ecozone<sup>+</sup> include the Mi'kmaq, Maliseet (of NB), Malécite (of QC), and Abenaki.

**Population:** Between 1971 and 2006, the human population increased from approximately 2.27 to 2.55 million (Figure 3), but has been generally stable since 1991. The majority of the population is found in its river valleys and along its low-lying coast.<sup>9, 10</sup> There has been a significant migration of people from rural to urban areas.<sup>11</sup>

Based on 2005 remote sensing data, forest is the predominant land cover type representing over 85% of the total area, followed by agricultural land at just over 10% (Figure 2).<sup>ii, 7</sup>

Water. See Ahern et al. 7 for further details.

<sup>&</sup>lt;sup>1</sup> Using remote sensing data, Ahern *et al.*<sup>7</sup> classified land cover into nine possible classes: Forest (tree crown density >10%), Shrubland (tree crown density <10%, shrub cover >40%), Grassland (tree or shrub cover on <10% of the land, herbaceous vegetation present), Agricultural Land (includes cropland and cropland/woodland); Low Vegetation and Barren (shrubs <40%), Fire Scars, Urban, Snow/ice/glacier (permanent ice or snow), and Inland

<sup>&</sup>lt;sup>ii</sup> Statistics for the amount of each land cover type may differ from statistics used in other parts of this report, for example the agricultural land section. This is due to a combination of the inventory method used, the scale of the data used, and different classification schemes for land cover types.

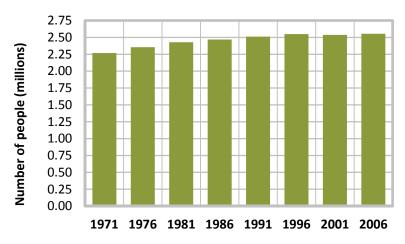
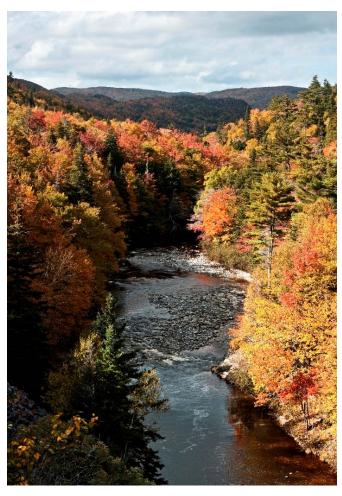


Figure 3. Human population of the Atlantic Maritime Ecozone $^{\scriptscriptstyle +}$ , 1971–2006. Source: Environment Canada, 2009 $^{^{12}}$ 



Cape Breton Highlands National Park, Nova Scotia © istockphoto.com / shaunl



Cavendish Beach, Prince Edward Island National Park © Parks Canada



Coastal marsh in Lord Selkirk Provincial Park, Prince Edward Island © iStock.com / Photawa



Peskawa Lake, Kejimkujik National Park, Nova Scotia © M. Crowley



Perce Village and Rock, Gaspé Peninsula, Quebec © iStock.com / onepony

# KEY FINDINGS AT A GLANCE: NATIONAL AND ECOZONE<sup>+</sup> LEVEL

Table 2 presents the national key findings from *Canadian Biodiversity: Ecosystem Status and Trends* 2010<sup>3</sup> together with a summary of the corresponding trends in the AME. Topic numbers refer to the national key findings in *Canadian Biodiversity: Ecosystem Status and Trends* 2010. Topics that are greyed out were identified as key findings at a national level but were either not relevant or not assessed for this ecozone<sup>+</sup> and do not appear in the body of this document. Evidence for the statements that appear in this table is found in the subsequent text organized by key finding. For many topics, additional supporting information can be found in the *Atlantic Maritime Ecozone*<sup>+</sup> *Status and Trends Assessment*.<sup>4</sup> See the Preface on page i.

Table 2. Key findings overview.

Themes and topics	Key findings: NATIONAL	Key findings: ATLANTIC MARITIME ECOZONE <sup>+</sup>
THEME: BIOMES		
1. Forests	At a national level, the extent of forests has changed little since 1990; at a regional level, loss of forest extent is significant in some places. The structure of some Canadian forests, including species composition, age classes, and size of intact patches of forest, has changed over longer time frames.	Forests cover approximately 80% of the AME. Historically, land clearing for agriculture and urban areas has reduced the extent of forests, including conversion of 70% of PEI's forests. No significant change in the extent of forests was found between 1985 and 2005. Although observed trends varied among regions, there was a shift from older to younger stands. Forest composition has shifted to simpler, less diverse forests as a result of forest clearing and regrowth, forestry, and natural disturbances. Intact forests larger than 50 km² cover only 5% of the AME.
2. Grasslands	Native grasslands have been reduced to a fraction of their original extent. Although at a slower pace, declines continue in some areas. The health of many existing grasslands has also been compromised by a variety of stressors.	Not relevant

Themes and topics	Key findings: NATIONAL	Key findings: ATLANTIC MARITIME ECOZONE <sup>†</sup>
3. Wetlands	High loss of wetlands has occurred in southern Canada; loss and degradation continue due to a wide range of stressors. Some wetlands have been or are being restored.	Wetlands cover 3.5% of the AME. It is not possible to determine trends in wetland extent over the entire AME, however, approximately 16–18% of freshwater wetlands in Nova Scotia were lost between European settlement and 1998. Many wetlands in the AME remain under continued threat of loss and degradation due to industrial and urban development, port expansion, cottage subdivisions, and agriculture. However, each of the four provinces have wetland conservation or similar policies that have mitigated the impacts of development projects and land-use decisions to some degree.
4. Lakes and rivers	Trends over the past 40 years influencing biodiversity in lakes and rivers include seasonal changes in magnitude of stream flows, increases in river and lake temperatures, decreases in lake levels, and habitat loss and fragmentation.	Between 1961–1982 and 1983–2003, changes in river flows included earlier onsets of spring freshet and decreased summer flows (low period from August to October). From 1970 to 2005, both minimum and maximum flows decreased at a high proportion of sites and minimum flows occurred later in the year. Dams have led to the local and regional extirpations of several plant, fish, and shellfish species.
5. Coastal	Coastal ecosystems, such as estuaries, salt marshes, and mud flats, are believed to be healthy in less-developed coastal areas, although there are exceptions. In developed areas, extent and quality of coastal ecosystems are declining as a result of habitat modification, erosion, and sea-level rise.	Coastal habitats have suffered significant losses and degradation as a result of human activities including industrial and urban development and cottage subdivisions. Coastal wetland loss and fragmentation is one of the most severe cases of wetland loss in Canada with 65% loss since the beginning of European settlement. Significant losses in beach and dune ecosystems have also been documented. Sea-level rise and increases in the frequency and intensity of storm surge events may increase erosion and flooding. Some species dependent upon coastal habitats, such as shorebirds, have also declined.

Themes and topics	Key findings: NATIONAL	Key findings: ATLANTIC MARITIME ECOZONE <sup>†</sup>
6. Marine	Observed changes in marine biodiversity over the past 50 years have been driven by a combination of physical factors and human activities, such as oceanographic and climate variability and overexploitation. While certain marine mammals have recovered from past overharvesting, many commercial fisheries have not.	Not relevant. Marine ecozones <sup>+</sup> are assessed in other ESTR reports.
7. Ice across biomes	Declining extent and thickness of sea ice, warming and thawing of permafrost, accelerating loss of glacier mass, and shortening of lake-ice seasons are detected across Canada's biomes. Impacts, apparent now in some areas and likely to spread, include effects on species and food webs.	There is a lack of long-term records for river and lake ice break-up and freeze-up. Available information showed no clear trends. There was a non-significant tendency toward decreasing sea ice cover and in the length of the ice season on the St. Lawrence River from 1971 to 2005.
THEME: HUMAN/ECO	DSYSTEM INTERACTIONS	
8. Protected areas	Both the extent and representativeness of the protected areas network have increased in recent years. In many places, the area protected is well above the United Nations 10% target. It is below the target in highly developed areas and the oceans.	In 2009, almost 11,000 km² (5.3%) of the AME was protected, an increase from just over 3,000 km² (1.6%) in 1992. This includes 10,963 km² (4.9%) protected as IUCN categories I–IV.
9. Stewardship	Stewardship activity in Canada is increasing, both in number and types of initiatives and in participation rates. The overall effectiveness of these activities in conserving and improving biodiversity and ecosystem health has not been fully assessed.	Trends in stewardship activities were not assessed for the AME for this current assessment thus this key finding is not included in this summary.
10. Invasive non- native species	Invasive non-native species are a significant stressor on ecosystem functions, processes, and structure in terrestrial, freshwater, and marine environments. This impact is increasing as numbers of invasive non-native species continue to rise and their distributions continue to expand.	Because it has many ocean ports, the AME has been a point of entry for many invasive non-native species, which are a threat to native biodiversity. Introduced invasive species have altered wetlands and coastal ecosystems, while invasive non-native insects and diseases have impacted forest ecoystems. Trend data do not exist for many species.

Themes and topics	Key findings: NATIONAL	Key findings: ATLANTIC MARITIME ECOZONE <sup>+</sup>
11. Contaminants	Concentrations of legacy contaminants in terrestrial, freshwater, and marine systems have generally declined over the past 10 to 40 years. Concentrations of many emerging contaminants are increasing in wildlife; mercury is increasing in some wildlife in some areas.	While relevant, contaminant trends were not assessed so this key finding is not included in this summary.
12. Nutrient loading and algal blooms	Inputs of nutrients to both freshwater and marine systems, particularly in urban and agriculture-dominated landscapes, have led to algal blooms that may be a nuisance and/or may be harmful. Nutrient inputs have been increasing in some places and decreasing in others.	Croplands in the AME have high levels of residual soil nitrogen and levels have increased from 1981 to 2006. As a result, the potential for leaching of nitrate out of soils and into water is high. A corresponding increase in nitrates has been found in ground and surface water in PEI. From 2002 to 2008, 18 estuaries had recurring anoxic events. Risk of surface water contamination from soil phosphorus has gradually shifted from lower to higher risk classes since 1991, with a higher proportion of cropland exceeding threshold values. In part of the Quebec portion of the AME, the number of lakes and rivers with algal blooms increased from three to 16 lakes between2004 and 2008.
13. Acid deposition	Thresholds related to ecological impact of acid deposition, including acid rain, are exceeded in some areas, acidifying emissions are increasing in some areas, and biological recovery has not kept pace with emission reductions in other areas.	Parts of the AME are highly sensitive to acid. Levels of sulphate and nitrate deposition decreased substantially between 1990 and 2004. Nevertheless, from 1999 to 2003, atmospheric sulphur and nitrogen deposition exceeded the critical load in several areas. The AME is North America's most heavily affected region in terms of the percentage of fish habitat lost due to acid rain with many rivers in Nova Scotia no longer supporting salmon.
14. Climate change	Rising temperatures across Canada, along with changes in other climatic variables over the past 50 years, have had both direct and indirect impacts on biodiversity in terrestrial, freshwater, and marine systems.	Between 1950 and 2007, summer temperatures increased by 1.1°C and precipitation increased 18.6% in fall. The number of days with precipitation increased in spring, summer, and fall. Future climate predictions include increased air and water temperatures, a longer growth season, and less sea ice cover in the Gulf of St. Lawrence as well as changes in storm intensity and frequency, and forest composition.

Themes and topics	Key findings: NATIONAL	Key findings: ATLANTIC MARITIME ECOZONE <sup>+</sup>
15. Ecosystem services	Canada is well endowed with a natural environment that provides ecosystem services upon which our quality of life depends. In some areas where stressors have impaired ecosystem function, the cost of maintaining ecosystem services is high and deterioration in quantity, quality, and access to ecosystem services is evident.	Important ecosystem goods and services in the AME include forest products, water, food production, fishing, hunting, wastewater assimilation, and tourism. The combined estimated value of ecosystem goods and services for the Atlantic provinces (excluding the Quebec portion of the AME) is over \$4.7 billion.
THEME: HABITAT, WIL	DLIFE, AND ECOSYSTEM PROCESSES	
16. Agricultural landscapes as habitat	The potential capacity of agricultural landscapes to support wildlife in Canada has declined over the past 20 years, largely due to the intensification of agriculture and the loss of natural and semi-natural land cover.	Although wildlife habitat capacity on agricultural landscapes remained high, it declined significantly between 1986 and 2006 due to expansion of cropland onto cover types that are less supportive of wildlife. Cropland in the AME has some of the highest erosion risk on agricultural land in Canada due to intensive tillage and a climate that poses a high threat of water erosion in some areas. However, soil erosion risk declined from 1981 to 2006.
17. Species of special economic, cultural, or ecological interest	Many species of amphibians, fish, birds, and large mammals are of special economic, cultural, or ecological interest to Canadians. Some of these are declining in number and distribution, some are stable, and others are healthy or recovering.	The Atlantic-Gaspésie population of the woodland caribou was reassessed in 2002 and its status updated from Threatened to Endangered. Moose populations have declined while white-tailed deer populations have expanded. The Bay of Fundy populations of Atlantic salmon face imminent extinction. Trends in other populations of Atlantic salmon were mixed, although many have declined. All landbirds, except forest birds, declined from the 1970s to the 2000s, with the greatest declines in bird assemblages of grassland habitats and other open habitats.
18. Primary productivity	Primary productivity has increased on more than 20% of the vegetated land area of Canada over the past 20 years, as well as in some freshwater systems. The magnitude and timing of primary productivity are changing throughout the marine system.	From 1985 to 2006, primary productivity, as measured by the Normalized Difference Vegetation Index, increased for 33,408 km² (16.5%) of the AME and decreased for 720 km² (0.4%). Areas of increase were concentrated in mixed forests along the Gaspé Peninsula and on Cape Breton Island, and were possibly a result of commercial logging.

Themes and topics	Key findings: NATIONAL	Key findings: ATLANTIC MARITIME ECOZONE <sup>+</sup>		
19. Natural	The dynamics of natural disturbance regimes, such as	Natural disturbance regimes are highly altered. Though		
disturbance	fire and native insect outbreaks, are changing and this	historically important, fire is of lesser significance due to		
	is reshaping the landscape. The direction and degree of	early detection and active suppression. Extreme weather		
	change vary.	events and insect outbreaks are the primary disturbance		
		agents. The frequency and severity of tropical storms and		
		hurricanes has increased from 1900 to 2000. Also, trend data		
		from Charlottetown, PEI, suggests an increase in the severity		
		and frequency of storm surge events. Spruce budworm is the		
		most influential insect. Although there is no consensus on		
		whether spruce budworm outbreaks are increasing in		
		frequency or severity, human activities are a factor.		
20. Food webs	Fundamental changes in relationships among species	Due to a variety of human-caused pressures, larger		
	have been observed in marine, freshwater, and	mammalian predators have been under continuous pressure		
	terrestrial environments. The loss or reduction of	in the AME. Top predators including wolves have been		
	important components of food webs has greatly	extirpated and other predators, such as American martens,		
	altered some ecosystems.	black bears, and lynx, have been lost from certain regions.		
		Coyotes have expanded their range into the AME.		

THEME: SCIENCE/POLICY INTERFACE									
21. Biodiversity monitoring, research, information management, and reporting	Long-term, standardized, spatially complete, and readily accessible monitoring information, complemented by ecosystem research, provides the most useful findings for policy-relevant assessments of status and trends. The lack of this type of information in many areas has hindered development of this assessment.	Long-term status and trend data are unavailable for many ecosystem components, particularly wetlands, changes in trophic structure, nonvascular plants, and invertebrates. Comprehensive data covering the entire AME is also rare; some results are available from case studies.							
22. Rapid change and thresholds	Growing understanding of rapid and unexpected changes, interactions, and thresholds, especially in relation to climate change, points to a need for policy that responds and adapts quickly to signals of environmental change in order to avert major and irreversible biodiversity losses.	As a result of the low capacity for buffering against acid in much the AME, thresholds for acid deposition were exceeded and Atlantic salmon populations declined. Rivers have not recovered with reductions in acid deposition. Due to impacts from forest management practices and fragmentation, forests have reduced capacity to support native species, such as large mammals. Climate change has and will continue to interact with other stressors, such as coastal erosion, invasive non-native species, and insect outbreaks, accelerating ecosystem damage.							

# **THEME: BIOMES**

Key finding 1 Theme Biomes

### **Forests**

### National key finding

At a national level, the extent of forests has changed little since 1990; at a regional level, loss of forest extent is significant in some places. The structure of some Canadian forests, including species composition, age classes, and size of intact patches of forest, has changed over longer time frames.

Forest is the predominant land cover in the AME, although total estimates have varied depending on the methods used. Using a mix of remote and ground-based sampling, the Canadian National Forest Inventory found that forests comprised approximately 77% of the area of the AME in 2001, consisting of 44% coniferous, 33% mixed, and 21% broadleaf forest. Based on 2005 remote sensing data, Ahern *et al.*7 estimated forest cover at over 85%. Differences between the two estimates reflect different methodologies and definitions of forest rather than a change in the area of forest in the AME<sup>iii</sup>.

Land clearing for agriculture and urban areas reduced the extent of post-settlement forests in certain areas, yet the AME remains well forested overall. By the beginning of the 20<sup>th</sup> century, 70% of the forest had been cleared for agriculture in PEI.<sup>14</sup> Agricultural land also replaced forest in much of Nova Scotia's Annapolis and New Brunswick's Saint John river valleys. However, Ahern *et al.*<sup>7</sup> found no significant change in the extent of forest cover between 1985 and 2005 based on remote sensing data. An analysis of forest density found that, other than in the Annapolis River Valley, the Saint John River Valley, and on most of Prince Edward Island, forest density was high for most of the AME. Over 30% of the 1 km² cells within the AME were more than 90% forested and another 20% were more than 80% forested.<sup>7</sup>

Age class distribution and composition of forests have changed through time, however, drawing general conclusions for the AME as a whole was difficult because there were no long-term data sets that covered the entire ecozone<sup>+</sup>. In general, the successional stage and age distribution of the forest shifted from old-growth to younger age classes.<sup>15</sup> Over the past several decades, forests have also become simplified in species and ecosystem diversity as a result of forest clearing and regrowth and natural disturbances.<sup>16</sup>

For more than 300 years the economy of the region has been dependent on forests to supply a diversity of products and services.<sup>17</sup>

-

The Canadian National Forest Inventory<sup>10, 13</sup> used inventory data from provincial, territorial, and other forest management agencies as well as remote sensing data to estimate forest cover. In contrast, Ahern *et al.*<sup>7</sup> (see Figure 2) is based solely on remote sensing data (and defines Forest as areas with tree crown density >10%).

There are three forest regions in the AME (Figure 4):18

- 1. The Acadian Forest Region, which extends into the northeastern United States, includes all of Nova Scotia and Prince Edward Island, and all but the northwestern corner of New Brunswick. It occupies an area of 122,000 km², is entirely within the AME in Canada, and represents 44% of the area of the AME.<sup>19, 20</sup> The region is transitional between the mostly deciduous forests of the south and west and the boreal coniferous forests of the north, and includes components of both.<sup>16</sup>
- 2. The Great Lakes–St. Lawrence Forest Region is predominantly a closed, mixed coniferous-deciduous forest. It is strongly influenced by the warm summers of the maritime climate that allow hardwoods to thrive. The forest region extends inland from the Great Lakes and St. Lawrence River to southeastern Manitoba, excluding the area north of Lake Superior. In the AME, this region occupies the northeast corner of New Brunswick, part of the Gaspé Peninsula, and the southern shore of the St. Lawrence River.
- 3. The Boreal Forest Region extends in a continuous belt from Newfoundland and Labrador west to the Rocky Mountains and north to Alaska. In the AME, it stretches from the northwestern tip of New Brunswick into the Gaspé Peninsula. This forest region is mostly coniferous, with black spruce (*Picea mariana*) and balsam fir (*Abies balsamea*) as principal species, but also includes some deciduous trees, such as white birch (*Betula papyrifera*) and trembling aspen (*Populus tremuloides*).

Quebec uses major bioclimate domains and subdomains to classify its forests and four of these include parts of the AME: sugar maple–basswood (east), sugar maple–yellow birch (east), balsam fir–yellow birch (east), and balsam fir–white birch (east) (Figure 5). These subdomains also include parts of the Boreal Shield and Mixedwood Plains ecozones<sup>+</sup>. Because the sugar maple—basswood (east) mostly includes land outside the AME, it is not included here. The Quebec forest subdomains in the AME overlap with the Great Lakes–St. Lawrence and Boreal forest regions defined above.

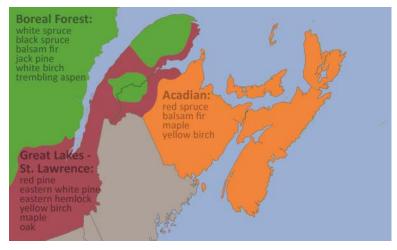


Figure 4. Forest regions and the principal tree species within each region. Source: Natural Resources Canada, 2007<sup>21</sup>

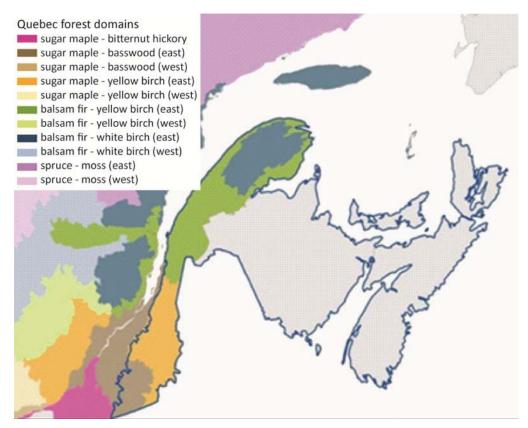


Figure 5. Quebec forest domains.

Source: Ministère des Ressources naturelles et Faune, 2005<sup>22</sup>

## Forest age structure

It has been estimated that as much as 50% of the forest in the AME may have been dominated by late-successional, old-growth forest types before European settlement.<sup>15</sup> In 2003, only about 1–5% of forests were estimated to be older than 100 years, and ground surveys suggest that far less than this had true old-growth forest characteristics.<sup>15</sup> A high proportion of the forests fell within the young age classes in 1999 (Figure 6), reflecting re-growth after forest harvesting.<sup>10</sup>

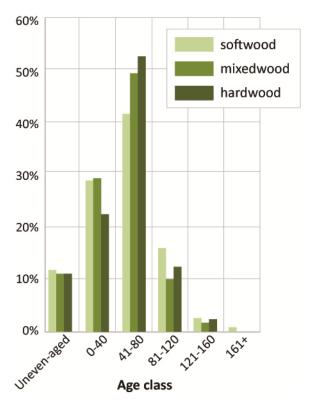


Figure 6. Age-class distribution by forest type on stocked forest land as a percentage of the total stocked forest area in the Atlantic Maritime  $Ecozone^+$ , 1999.

Source: adapted from Canadian Council of Forest Ministers, 2005<sup>10</sup>

#### **Acadian forest**

Younger forests increased and older forests declined between 1958 and 2003 in Nova Scotia (Figure 7).<sup>23</sup> The youngest age class (less than 20 years) increased as a proportion of total forest cover, from 3.8% in the early 1970s to 23.9% in the 1997–2003 inventory. Forests greater than 101 years of age decreased from 8.7% in 1958 to 0.3% in the 1997–2003 inventory and forests between 81 and 100 years decreased from 16.4 to 1.2%.

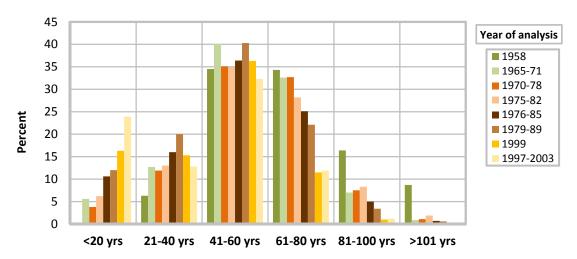


Figure 7. Percent of total forest area in each age class in Nova Scotia, 1958–2003.

Time periods are ranges and vary as Pannozza and Coleman (2008) compiled data from several sources: Forest Resources of Nova Scotia (1958); Nova Scotia Forest Inventory provincial summary (1965–1971, 1970–1978, 1975–1982, 1976–1985, 1979–1989); Department of Natural Resources GIS 1995 inventory data (1999); and Department of Natural Resources GIS unpublished inventory data (1997–2003). Source: adapted from Pannozza and Coleman, 2008<sup>23</sup>

### Forest subdomains in Quebec portion of AME

Trends from the 1970s to the 1990s showed a gain in balsam fir domain forest stands in a mature developmental stage (Figure 8). Within this 30-year period, 19% of the balsam fir—yellow birch forest sub-domain became mature, while 23% was lost from the young category.<sup>24</sup> In the balsam fir—white birch sub-domain, mature forest stands and regenerated stands remained stable (2% increase and 1% decrease respectively) over the same time period, while young forest stands decreased by 5% and regenerating forest stands increased by 3%. For the sugar maple—yellow birch subdomain, young forest stands decreased by 3% and regenerated stands increased by 6%.<sup>24</sup> Over the 30-year period, the proportion of mature stands did not change. These data included areas outside the AME.

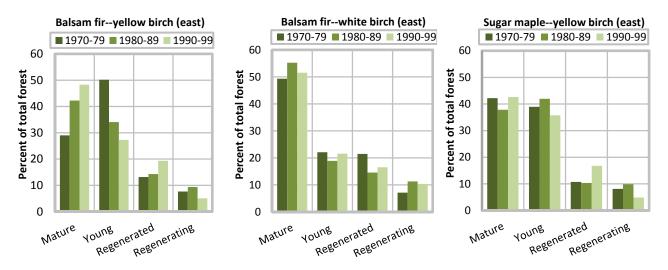


Figure 8. Proportion of forest at each major developmental stage in Quebec's subdomains that occur in the Atlantic Maritime  $Ecozone^{+}$ , 1970s, 1980s, and 1990s.

Development stages are based on stand height and growth in volume: regenerating = disturbed stands <2 m in height; regenerated = disturbed stands 2–7 m in height; young = stands >7 m with increasing mean annual growth (volume); mature = stands >7 m with decreasing mean annual growth (volume). Data included area outside the AME.

Source: Ministère des Ressources naturelles et Faune, 2009, Statistiques forestières, unpublished data; updated from Ministère des Ressources naturelles, 2002<sup>25</sup>

## Forest composition

In many parts of the AME, the forest has been simplified both in species and ecosystem diversity over the past several decades. This was primarily a result of forest clearing for agriculture and subsequent abandonment, timber removal of selected species, and clear-cutting as well as natural disturbance.<sup>16</sup>

#### **Acadian forest**

As older forests were replaced by relatively young, often even-aged, early successional forest types, the abundance and age of late-successional species such as sugar maple (*Acer saccharum*), red spruce (*Picea rubens*), eastern hemlock (*Tsuga canadensis*), red oak (*Quercus rubra*), yellow birch (*Betula alleghaniensis*), American beech (*Fagus grandifolia*), and eastern white cedar (*Thuja occidentalis*) declined.<sup>16, 26</sup> Younger forests have higher frequencies of balsam fir, red maple (*Acer rubrum*), white spruce (*Picea glauca*), white birch, and trembling aspen.<sup>16, 27</sup> Similar changes are occurring in other eastern forests where species composition was altered by logging and land clearing throughout the twentieth century.<sup>28</sup>

A case study in Kings County, NB, compared forest species composition in 1800 and 1993. Species distribution in 1800 was more even than in 1993 (Figure 9). The study showed that cedar was likely as common as balsam fir in the early 1800s, but by the 1990s, balsam fir was four times as common as cedar. The spruce genus increased in frequency, as did poplar, and

white pine remained stable, but the rest of the other tree genera were more common 200 years ago than today. Cedar, hemlock, ash, beech, and larch declined over the time period. Balsam fir and the spruces comprised about 50% of the forest in 1993, while 200 years ago, they accounted for only 25%. <sup>16</sup>

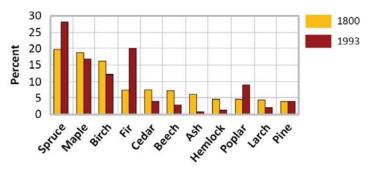


Figure 9. Estimated frequency of major forest tree genera in Kings County, NB, 1800 and 1993. Source: Loo and Ives, 2003<sup>16</sup>

### Forest subdomains in Quebec portion of AME

From the 1970s to 1990s, conifers, particularly balsam fir, in the balsam fir subdomains declined,<sup>25</sup> while mixedwood stands increased. In the balsam fir–yellow birch and balsam fir–white birch subdomains, conifer proportions decreased by 8 and 16%, respectively. Conifers in the sugar maple–yellow birch subdomain declined, while mixedwood stands increased (Figure 10). These data include public and private forests, and include all the area of the subdomains, including some area outside the AME.

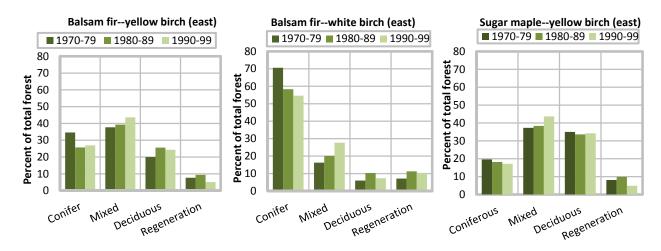


Figure 10. Proportion of total area covered by different forest cover types in the Quebec subdomains that occur in the Atlantic Maritime Ecozone $^+$ , 1970s, 1980s, and 1990s.

Forest cover types are based on stand height and composition; regenerating = <2 m in height; deciduous = >75% deciduous; mixed = 25-75% deciduous; coniferous = >75% coniferous. Data included area outside the AME.

Source: Ministère des Ressources naturelles et Faune, 2009, Statistiques forestières, unpublished data; updated from Ministère des Ressources naturelles, 2002<sup>25</sup>

Differences in land cover types were also measured between 1993 and 2001 for the Appalachian Ecoregion which overlaps with parts of the area above. Jobin *et al.*<sup>29</sup> reported that the abundance of mixedwood forest declined by over 12% while coniferous stands increased by 7% over this time period (Figure 11).

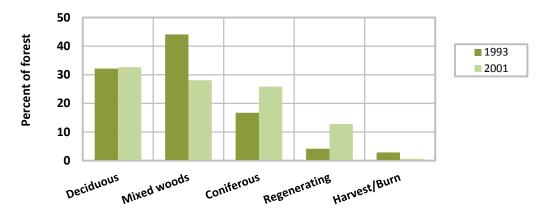


Figure 11. Change in forest types for the Appalachian Ecoregion in southern Quebec between 1993 and 2001.

Source: Jobin et al., 2007<sup>29</sup>

## **Fragmentation**

Fragmentation reduces habitat connectivity, increases edge density, and increases the isolation of remnant habitat patches. In contrast to more remote, less populated ecozones<sup>+</sup>, remaining natural ecosystems of the AME are highly fragmented. Only 5% of the AME is covered by intact fragments of natural ecosystems (primarily forests) larger than 50 km² (Figure 12).<sup>19</sup>

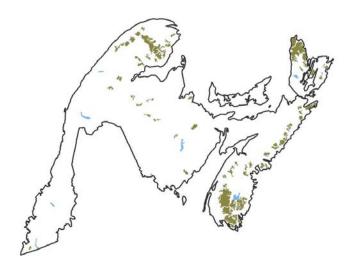


Figure 12. Intact landscape fragments larger than 50 km² in the Atlantic Maritime Ecozone†, 2003. A landscape fragment is a contiguous mosaic of various ecosystems, naturally occurring and essentially undisturbed by significant human influence.

Source: adapted from Lees et al., 2006<sup>19</sup>

### Forest birds

Changes in the age structure of the forest, with increasing early-successional stands and decreasing contiguous mature stands, and the replacement of some hardwood stands with softwood plantations and agricultural land have resulted in changes in the bird community. Overall, forest bird populations have been generally stable but with a tendancy to decline, especially since 2000 (Figure 13). There have been large declines for several species, while others have stable or increasing populations. For example, the Canada warbler (*Cardellina canadensis*), a species recently assessed as Threatened by the Committee on the Endangered Status of Wildlife in Canada (COSEWIC), has declined by 80% in the AME since the 1970s. Although the primary cause of its decline is unclear, research has shown this species is sensitive to forest fragmentation and human disturbance. Populations may have been affected on both the breeding and wintering grounds by habitat loss and degradation. The decline in spruce budworm (*Choristoneura fumiferana*) abundance may also have reduced an important food source for Canada warbler. <sup>32, 33</sup> The boreal chickadee (*Poecile hudsonicus*) has also declined markedly in this region and throughout its range. <sup>34, 35</sup>

The AME region in Canada and similar neighbouring areas in the United States support over 90% of the world's breeding population of Bicknell's thrush (*Catharus bicknelli*), one of the rarest songbirds in North America and listed as Threatened in Canada.<sup>36</sup> This bird lives in the high-elevation coniferous forests and is particularly susceptible to climate change, which may result in shifts in high-elevation breeding zones. Other threats incude habitat loss and degradation on both the breeding and wintering grounds, squirrel predation at nests, and environmental contaminants.<sup>37-39</sup> Surveys over the last several years indicate this species has undergone considerable annual decline.<sup>35, 40, 41</sup>

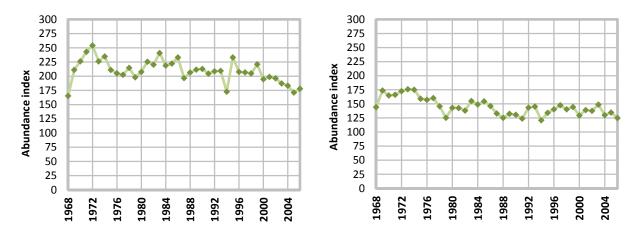


Figure 13. Annual indices of population change for birds of forest habitat (left) and shrub-early successional habitat (right), 1968–2006.

Shrub/early successional assemblage includes shrubland, old field, and mid-successional stage habitat from grassland to forest.

Source: Downes et al., 2011<sup>42</sup>using data from the Breeding Bird Survey<sup>43</sup>

A large portion of forested land in the ecozone<sup>+</sup> is in early successional stages. The overall slightly negative trends in the indices of population change for birds inhabiting forest and shrub-early successional habitat types (Figure 13) were influenced by the strong declines in abundant species such as the white-throated sparrow (*Zonotrichia albicollis*) and song sparrow (*Melospiza melodia*). However, declines in these species have been largely balanced by increases in several generalist species, such as Nashville warbler (*Oreothlypis ruficapilla*), yellow warbler (*Setophaga petechia*), and chestnut-sided warbler (*Setophaga pensylvanica*), which utilize and have benefited from increases in shrub-early successional forest habitat.<sup>44</sup>

## Cumulative human impact

The organization Two Countries One Forest quantified the human footprint on terrestrial ecosystems of the Appalachian/Acadian Ecoregion (which includes the AME) by integrating four categories of human influence: settlement, access, land use, and electrical power infrastructure (Figure 14).<sup>45</sup> In 2008, the greatest human impacts were primarily on coastlines, valleys, and other low-lying areas, reflecting the historical pattern of settlement. Only 0.2% of

the ecoregion has a human footprint score of 0, indicating no human transformation of the landscape. More than 90% of the ecoregion has a low human footprint (score of less than 50). Large areas are classified as having low impact; however, they tend to be separated by areas with high levels of human activity, thus fragmenting the region.<sup>45</sup> Only 5% of the total landscape is in intact natural fragments of larger than 50 km<sup>2</sup>.<sup>19</sup>

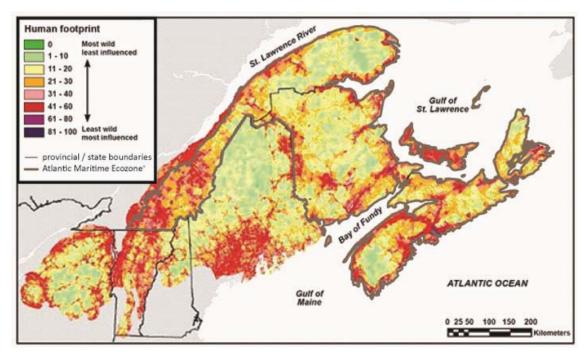


Figure 14. The human footprint of the Northern Appalachian/Acadian Ecoregion, 2008. Source: Trombulak et al., 2008<sup>45</sup>

Key finding 3 Theme Biomes

### Wetlands

#### **National key finding**

High loss of wetlands has occurred in southern Canada; loss and degradation continue due to a wide range of stressors. Some wetlands have been or are being restored.

There has been extensive wetland loss in the AME, particularly in coastal areas, where shoreline development is a continuing threat (see Coastal wetlands section on page 29). Although three of the four provinces in the AME have had wetland inventories completed since the 1980s, wetland mapping and assessment methodologies have changed, making it difficult to determine the amount of change over time. <sup>10, 46</sup> According to data from Canada's Forest Inventory (CanFI), <sup>13</sup> in 2001, wetlands covered approximately 3.5% of the land area of the ecozone<sup>+</sup> and approximately 35% of those were treed. <sup>10</sup> Although estimates of freshwater

wetland losses are not available for the AME as a whole, approximately 16 to 18% of freshwater wetlands in Nova Scotia had been developed or converted to other ecosystem types between European settlement and 1998.<sup>47</sup> Coastal wetland loss in Nova Scotia has been estimated at 65%.<sup>48</sup>

Many wetlands in the AME remain under continued threat of loss and degradation due to industrial and urban development, port expansion, cottage subdivisions, and agriculture. However, each of the four provinces have wetland conservation or similar policies that have mitigated the impacts of development projects and land-use decisions to some degree. Bogs are being impacted by commercial peat moss extraction and cranberry production.

## Waterfowl

Trends for selected breeding waterfowl species show either stable or increasing populations since 1993 (Table 3).<sup>50</sup> The American black duck (*Anas rubripes*), the most abundant duck in the AME, has been the focus of special conservation effort because the wintering population in the United States decreased by almost 50% between 1955 and 1985.<sup>51, 52</sup> In the AME, from 1993 to 2006, black duck populations were stable (Table 3).<sup>50</sup> Logging, hydroelectric development, transmission line construction, agriculture, urbanization, and industrial development threaten breeding and staging habitats.<sup>51</sup> In addition, it is likely that the species has had to compete for habitat with a growing mallard (*Anas platyrhynchos*) population.<sup>53</sup> Some evidence shows that habitat availability and quality may not be limiting, however,<sup>53</sup> and recent increases and stabilization of the black duck may reflect increased hunting restrictions in Canada and the United States.<sup>52</sup> Black ducks are also closely related to mallards and the two species interbreed regularly, which may represent an additional conservation concern for the species.<sup>54-56</sup>

Table 3. Abundance trends for selected breeding waterfowl species in the Atlantic Maritime Ecozone<sup>+</sup>, 1990s−2000s.

Cassina	Nesting habitat	Trend (%/yr)	Р	Annual Index (in thousands)		
Species				1990s	2000s	% change
Mallard (Anas platyrhynchos)	Ground	30.1	*	2.3	4.6	98.1
American black duck (Anas rubripes)	Ground	2.2		57.7	63.7	10.5
Green-winged teal (Anas crecca)	Ground	5.9	n	8.4	11.7	38
Ring-necked duck (Aythya collaris)	Overwater	6.5	*	21.2	32.3	52.2
Canada goose (Branta canadensis)	Ground	22.5	*	1.1	3.6	244.3

P is the statistical significance: \* indicates P<0.05; n indicates 0.05<P<0.1; no value indicates not significant

For a description of how species were selected and data methodology, see Fast *et al.*, 2011.

Source: Fast et al., 2011<sup>50</sup>

Key finding 4 Theme Biomes

## Lakes and rivers

#### **National key finding**

Trends over the past 40 years influencing biodiversity in lakes and rivers include seasonal changes in magnitude of stream flows, increases in river and lake temperatures, decreases in lake levels, and habitat loss and fragmentation.

Of the 1,792 lakes in the Atlantic Provinces, 98% are less than 99 km² in size.<sup>57</sup> The largest lake in the AME is Grand Lake, NB. The Saint John River in New Brunswick is the largest river system in the AME.<sup>58</sup> Seven rivers/river systems within the AME are classified as Canadian Heritage Rivers: the Saint John, St. Croix, and Upper Restigouche rivers in New Brunswick; the Shelburne River and Margaree-Lake Ainslie river system in Nova Scotia; and the Hillsborough River and Three Rivers (Cardigan, Brudenell, and Montague/Valleyfield) river system on Prince Edward Island. <sup>58, 59</sup> Runoff increases significantly from west to east, varying from 60 cm annually in the western part of the AME to 200 cm along the Atlantic coast.<sup>58-61</sup>

Lakes and rivers in the AME support diverse aquatic communities including species at risk such as the Atlantic salmon (*Salmo salar*), striped bass (*Morone saxatilis*), Atlantic whitefish (*Coregonus huntsmani*), American eel (*Anguilla rostrata*), wood turtle (*Glyptemys insculpta*), Blanding's turtle (*Emydoidea blandingii*), dwarf wedgemussel (*Alasmidonta heterodon*), yellow lampmussel (*Lampsilis cariosa*), skillet clubtail (*Gomphus ventricosus*), cobblestone tiger beetle (*Cicindela marginipennis*), and several coastal plain flora.

# Streamflow in natural rivers

Two analyses of streamflow in rivers with minimal flow control or impact upstream over the past 40 years were conducted for ESTR. Cannon *et al.*<sup>62</sup> looked at seasonal trends in streamflow at sites across Canada between two periods, 1961–1982 and 1983–2003. To facilitate the analysis of trends nationally, sites were organized into six groups with similar intra-seasonal patterns of flow (hydrology groups). Across sites in the AME, changes in flows between 1961–1982 and 1983–2003 included earlier onsets of spring freshet and decreased summer flows (summer flow period from August to October).<sup>62</sup> Monk and Baird<sup>63</sup> found that minimum and maximum flow variables decreased at a high proportion of sites from 1970 to 2005 and the annual 1-day minimum flow occurred later in the year. Although the rise rate decreased significantly at 32% of the sites and the fall rate increased at 29%, no overall trend was found in the variability of annual runoff.<sup>63</sup> Figure 15 summarizes the number and direction of significant trends in streamflow variables for the 34 stations analyzed by Monk and Baird<sup>63</sup> and Figure 16 shows the results of the Cannon *et al.*<sup>62</sup> analysis of seasonal trends at representative sites.

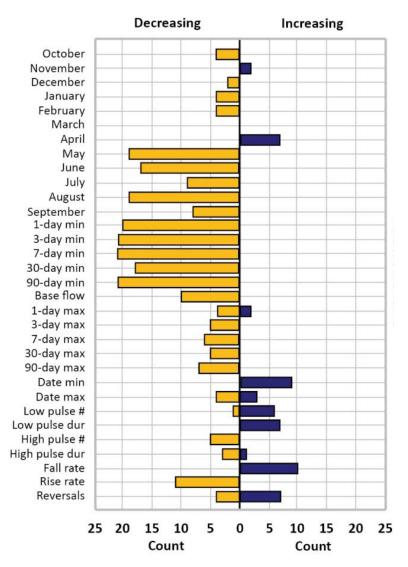


Figure 15. Summary of the total number of sites displaying increasing and decreasing trends in various streamflow variables in the Atlantic Maritime Ecozone<sup>+</sup>, 1970–2005.

Based on 34 gauging sites. Only sites with significant trends (p<0.1) are shown.

Source: Monk and Baird, 2014<sup>63</sup>

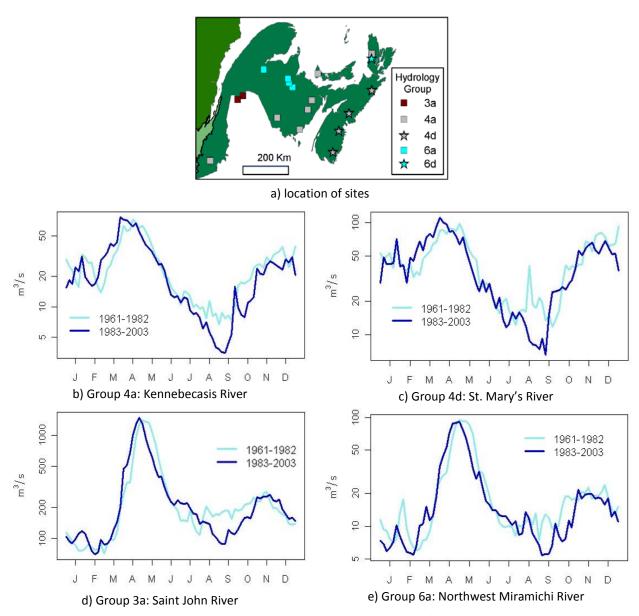


Figure 16. Changes in streamflow between 1961–1982 and 1983–2003 for representative sites of each hydrology group in the Atlantic Maritime Ecozone $^+$ .

Hydrology groups represent clusters of rivers showing similar hydrologic responses to variations in climate. For information on the specific hydrology groups mentioned above, see Cannon et al. 2011. Source: Cannon et al., 2011

#### Water control structures

Although water control structures are less common in the AME than in some ecozones<sup>+</sup>, their impacts are often greater because of the coastal nature of the ecozone<sup>+</sup> and large number of watersheds with limited numbers of natural lakes. Ecological impacts include local and regional species extirpations and habitat loss and alteration. Impacts of these structures on lakes and rivers include altering natural water level fluctuations, peak flows, seasonal flooding, and

natural disturbance regimes, as well as decreasing water quality.<sup>64</sup> A total of 74 large dams (greater than 10 m in height) have been constructed in the AME, although few since the 1970s.

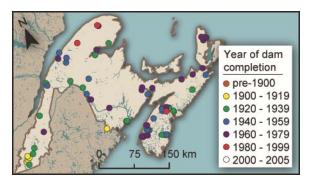


Figure 17. Spatial distribution of dams greater than 10 m in height within the Atlantic Maritime Ecozone<sup>+</sup>, grouped by year of completion between 1830 and 2005.

Source: Canadian Dam Association, 2003<sup>65</sup>

Examples of the impacts of dams on biodiversity in the AME include:

- extirpation of three plants, Canadian honewort (*Cryptotaenia canadensis*), prairie goldenrod [*Oligoneuron album* (syn. *Solidago ptarmicoides*)], and American bittersweet (*Celastrus scandens*) from the Saint John River, NB, due to flooding from hydroelectric dams;<sup>66</sup>
- local extirpation of two species at risk, Plymouth gentian (*Sabatia kennedyana*) and pink coreopsis (*Coreopsis rosea*), from at least two lakes in the Tusket River system in extreme southwestern Nova Scotia;<sup>67, 68</sup>
- extirpation of one of only two populations of the endangered Atlantic whitefish (*Coregonus huntsmani*) in Nova Scotia as a result of the damming the Tusket River in 1929;69 and
- extirpation of the dwarf wedgemussel (*Alasmidonta heterodon*), a species that was
  restricted to the AME in Canada, likely as a result of the loss of its fish host due to
  construction of a causeway over the tidal portion of the Petitcodiac River, NB, in 1967–
  1968.<sup>70,71</sup>

Key finding 5 Theme Biomes

## Coastal

#### **National key finding**

Coastal ecosystems, such as estuaries, salt marshes, and mud flats, are believed to be healthy in less-developed coastal areas, although there are exceptions. In developed areas, extent and quality of coastal ecosystems are declining as a result of habitat modification, erosion, and sea-level rise.

The coast is a defining feature of the AME. Coastal features include bays, coves, harbours, inlets, passages, channels, basins, points, heads, promontories, islands, capes, beaches, barrens, estuaries, and salt marshes.<sup>72</sup> There is some information on the status of coastal features in the AME, however, trends in this biome are not well known. Where data exist, they are not comprehensive and not necessarily representative of the entire AME. Given these caveats, available data suggest that coastal wetlands, beaches, and dunes declined and that stressors, such as development (industrial, urban, and cottage development),<sup>73</sup> recreation, sea-level rise, and storm surges increased. Climate change will increase impacts to these coastal habitats.<sup>73</sup> Some coastal dependent species, such as certain shorebirds, also declined.

## Coastal wetlands

Although coastal wetlands and shores cover less than 1% of the AME, they are one of the most important habitat types for maintaining native biodiversity. Loss and fragmentation of this ecosystem type in the AME is one of the most severe cases of wetland loss in Canada.<sup>27</sup> As already mentioned in the Wetlands section (page 24), an estimated 65% of the area covered by coastal marshes has been lost since European settlement.<sup>48</sup> Wetland loss began over 300 years ago when Acadians began draining salt marshes for agriculture. Since 1900, many coastal wetlands have been drained, flooded, and/or filled in for urban, industrial, or agricultural purposes and coastal developments, particularly cottage subdivisions.<sup>74</sup>

Hanson *et al.*<sup>75</sup> quantified change in the extent of salt marshes in two undeveloped (Cape Jourimain and Shemogue) and three developed (Aboiteau, Shediac, and Cocagne) sites along the Northumberland Strait in southeastern New Brunswick between 1944 and 2001. Salt marshes declined at all five sites over the study period from a combination of development and climatic variables (Figure 18).

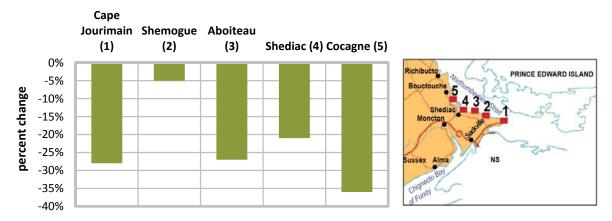


Figure 18. Decline in area of vegetated salt marsh in five locations in southeastern New Brunswick between 1944 and 2001.

Study sites Cape Jourimain (1) and Shemogue (2) are undeveloped areas. The other three sites (3–5) are largely residential.

Source: adapted from Hanson et al., 2006<sup>75</sup>

Coastal wetlands continue to be degraded as a result of terrestrial runoff and sedimentation, the restriction of tidal water movement due to barriers and culverts,<sup>73</sup> and the rise in sea levels due to climate change. Industrial and commercial development, as well as some agricultural practices, are among the principal threats to estuarine ecosystems.<sup>76</sup> Continued sea-level rise will result in additional negative impacts on the coast<sup>73</sup> (see Sea-level rise and coastal erosion section on page 36).

#### **Eelgrass**

Eelgrass meadows are among the most productive ecosystems in the world,<sup>77</sup> and also among the most threatened.<sup>78</sup> Eelgrass (*Zostera marina*) is an important food for migrating and wintering waterfowl, and provides foraging areas for other birds.<sup>79,81</sup> Comprehensive trend data do not exist for eelgrass but compiling results from a number of mainly short-term studies (Table 4) suggests a general decline in eelgrass and some abrupt die-offs, along with some areas with stable to increasing trends.<sup>77,80</sup> Loss of eelgrass beds worldwide have been attributed to a range of natural and human-induced disturbances, including coastal erosion, hurricanes, sediment and nutrient loading (see Nutrient loading and algal blooms section on page 52), and various forms of mechanical disturbance.<sup>82</sup> Another factor in declines on the Atlantic coast is the spread of the invasive green crab (*Carcinus maenas*), which can uproot eelgrass plants.<sup>83</sup>

Table 4. Trends in eelgrass from studies in Nova Scotia and the Gulf of the St. Lawrence.

Location	Years	Eelgrass trends	
Lobster Bay, NS	1978–2000	Estimated losses of 30 and 44% in two areas <sup>84</sup>	
Antigonish Harbour, NS	2000–2001	Biomass decline of 95% followed by 50% decline in geese	
		and ducks that feed on the eelgrass <sup>85</sup>	
4 Nova Scotia inlets	1992–2002	Loss of 80% of total intertidal area occupied by eelgrass <sup>86</sup>	
13 southern Gulf of	2001–2002	Biomass decline of 40% <sup>87</sup>	
St. Lawrence estuaries			
Gulf of St. Lawrence in	various	Manicouagan Peninsula distribution expanded (1986 to	
Quebec		2004); generally also expanding or stable in other areas <sup>88</sup>	

## **Shorebirds**

Although the AME supports a number of breeding shorebird species, it is most important for migrant shorebirds. Coastal habitats, particularly those around the upper Bay of Fundy, are of critical importance as stopover and refueling areas, particularly for the smaller sandpipers.<sup>89-91</sup> The number of shorebirds passing through the Canadian Atlantic provinces declined since surveys were started in 1974 (Table 5),<sup>92-96</sup> with declines particularly pronounced in the 1990s.<sup>97</sup> The reasons for the declines are not fully understood. Although coastal habitats have changed in ways that can negatively affect shorebirds,<sup>98</sup> trends in at least some species likely reflect factors in other parts of the birds' migration ranges.<sup>98</sup>

Table 5. Trends in abundance of shorebirds migrating through coastal areas of the Atlantic Maritime  $Ecozone^+$ , 1974–2006.

Spaciac	Trend	Р	Abundance index			Change	
Species	(%/yr)		1970s	1980s	1990s	2000s	%
Red knot	-10.9	*	39.5	11.2	9.1	3.3	-97.5
(Calidris canutus)							
Least sandpiper	-6.6	*	80.7	22.2	9.8	11.6	-88.8
(Calidris minutilla)							
Lesser yellowlegs	-5.0	*	29.2	52.2	16.4	9.8	-80.6
(Tringa flavipes)							
Semipalmated sandpiper	-4.9		5170.9	4892	2623.7	3074.5	-80.0
(Calidris pusilla)							
Black-bellied plover	-3.0	*	51.0	43.1	23.0	26.7	-62.3
(Pluvialis squatarola)							
Dunlin	-2.8		26.3	28.6	11.4	15.5	-59.7
(Calidris alpina)							
Ruddy turnstone	-2.8	**	13.2	10.9	11.4	4.2	-59.7
(Arenaria interpres)							
Short-billed dowitcher	-2.7		292.8	281.7	39.6	141.0	-58.4
(Limnodromus griseus)							
Sanderling	-2.3		42.9	34.7	19.8	24.0	-52.5
(Calidris alba)							
Greater yellowlegs	-0.9		13.0	12.8	9.8	10.8	-25.1
(Tringa melanoleuca)							
Hudsonian godwit	-0.9		5.5	4.1	3.5	2.9	-25.1
(Limosa haemastica)							
Willet	-0.8		16.6	15.9	11.1	14.1	-22.6
(Tringa semipalmata)							
White-rumped sandpiper	-0.2		16.1	15.3	12.6	16.4	-6.2
(Calidris fuscicollis)							
Semipalmated plover	1.9		103.8	123.0	153.1	159.3	82.6
(Charadrius semipalmatus)							
Whimbrel	2.5		1.9	1.5	3.1	4.3	120.4
(Numenius phaeopus)							

P is the statistical significance: \*\* indicates P<0.01, \* indicates P<0.05, no value indicates not significant "Change" is the percent change in the average abundance index over the entire period calculated from the overall trend (%/yr).

Source: Gratto-Trevor et al., 2011<sup>97</sup>

Relatively few species of shorebirds breed in the AME, and only a small number in coastal areas. Of the four coastal breeding species in the AME for which trends can be determined from Breeding Bird Survey data (Table 6), only the trend for Wilson's snipe (*Gallinago delicata*) was significant, declining at 2.6%/yr (P<0.01).<sup>97</sup>

Table 6. Trends in abundance o	f coastal breeding shorebirds	s, 1970s to 2000s.

Species	Trend	P	Abundance index			Change	
Species	(%/yr)		1970s	1980s	1990s	2000s	%
Upland sandpiper (Bartramia longicauda)	-3.1		3	0.2	0.1	0.1	-70%
Spotted sandpiper (Actitis macularius)	-2.6		0.8	0.9	0.7	0.4	-64%
Willet (Tringa semipalmata)	-3.1		1.1	1	0.4	0.4	-71%
Wilson's snipe (Gallinago delicata)	-2.6	**	5.2	4.8	2.9	2.3	-64%

P is the statistical significance: \*\* indicates P<0.01, no value indicates not significant Change indicates the percent change in the average abundance index over the survey period (1968–2006) calculated from the overall trend (%/yr).

Source: Gratto-Trevor et al., 2011<sup>97</sup> using data from the Breeding Bird Survey<sup>43</sup>

#### Waterfowl

The AME has many coastal areas where large numbers of waterfowl traditionally congregate during the spring and fall migrations.<sup>99</sup> Many waterfowl also winter in this region, for example, Barrow's goldeneye (*Bucephala islandica*).<sup>100</sup> Recent milder winters with longer ice-free periods have resulted in larger wintering populations and potential increases in the residency times of waterfowl during migration.<sup>101</sup> Trends in breeding waterfowl are summarized in the Wetlands section on page 24.

# Sandy shores and sand dunes

Sandy shores and sand dunes are primarily located along the New Brunswick coast of Northumberland Strait, the Minas Basin, the north shore of PEI, and Îles de la Madeleine. Beaches and dunes are important habitat for many species of wildlife, providing food and habitat to shorebirds and other fauna, flora, and microorganisms.<sup>76</sup> They are threatened by development, sand mining, recreation, sea-level rise, and increased storm severity related to climate change (see Sea-level rise and coastal erosion section on page 36).

Data on trends in erosion and deposition rates for beach and dune habitat is limited. O'Carroll *et al.*<sup>102</sup> conducted a retrospective analysis of aerial photos to assess temporal changes in beach and dune habitat at five locations in southeastern New Brunswick between 1944 and 2001. They found that the amount of beach and dune habitat had declined in all sites, with a greater decline in beach than in dune in all five locations (Figure 19). Sand removal for aggregate production and the expansion of shoreline protection have also contributed to changes in these areas. The variety of changes observed illustrates that local accretion and erosion processes, storm events, and human activity have all been important factors in shaping coastal sand ecosystems.<sup>102</sup>

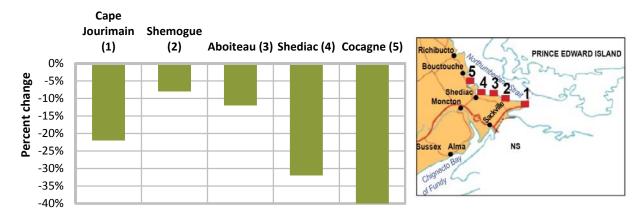


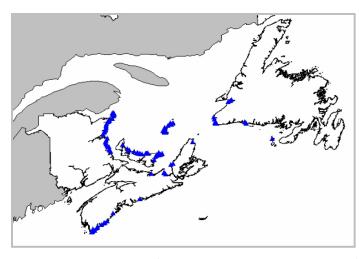
Figure 19. Decline in area of beach and dune habitat in five locations in southeastern New Brunswick between 1944 and 2001.

Study sites Cape Jourimain (1) and Shemogue (2) are undeveloped areas. The other three sites (3–5) are largely residential. For Shediac, the 32% decline was between 1944 and 1971 with little additional loss between 1971 and 2001.

Source: adapted from O'Carroll et al., 2006<sup>102</sup>

## **Piping plover**

The Atlantic population of piping plovers (*Charadrius melodus melodus*), listed as Endangered under Canada's *Species at Risk Act*, prefers early-successional habitat, such as barrier islands converted from sand spits by storm activity (Figure 20).<sup>73</sup> In 2002, the global piping plover breeding population was estimated at only 5,945 adults.<sup>103</sup> In the AME, 442 adults in 2001 and 435 adults in 2006 were detected at breeding sites. Despite active conservation programs there was been a 13% decline in the number of adults from 1991 to 2006 (Figure 20). There are several threats to piping plovers, with predation being one of the most important factors limiting populations across the North American breeding range. Current estimates in eastern Canada suggest that hatching success is less than 55%.<sup>104</sup> In addition, habitat loss and degradation are significant problems. Increased use of beaches and coastal development, including construction of cottages or homes, wharves, jetties, and erosion control structures can impact nesting beaches as well as brood-rearing and foraging habitat.<sup>105</sup> Impacts from climate change are another factor, including storm surges, which are becoming more frequent, and sea-level rise.<sup>106, 107</sup>



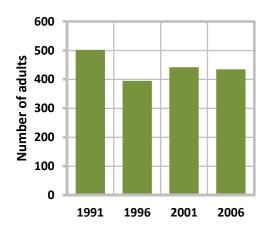


Figure 20. Distribution of 2006 piping plover nesting sites (left, map) and the number of piping plover adults counted during surveys (right, bar chart) in the Atlantic Maritime Ecozone $^+$ , 1991, 1996, 2001, and 2006.

Count data from International Piping Plover Censuses 1991–2006. Numbers reported reflect "high counts" and include all adults counted during all surveys at all sites (some sites surveyed multiple times). Source: map from Environment Canada, 2006;<sup>106</sup> data from Ferland and Haig, 2002<sup>103</sup> and Elliot-Smith et al., 2009<sup>108</sup>

# Coastal development

Since 1990, coastal areas of the AME have become more heavily populated. In New Brunswick, for example, the proportion of coastal subdivisions as a percentage of all subdivisions in the province increased 35% from 1990 to 1999. In Nova Scotia, increased urbanization led to population declines in many rural areas of the province, while populations increased along the coast. There was a dramatic increase in the rates of subdivision and lot registrations on coastal land through the 20th century (Figure 21). The subdivision are subdivision and lot registrations on coastal land through the 20th century (Figure 21).

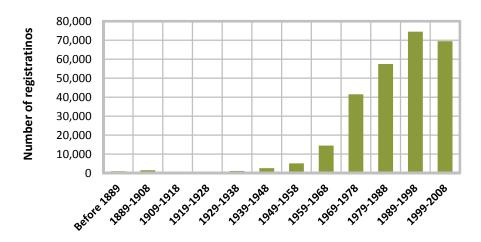


Figure 21. Trends in lot registration within two km of the Nova Scotia coastline by decade. Source: adapted from Nova Scotia Property Online Database by CBCL Limited, 2009<sup>76</sup>

## Sea-level rise and coastal erosion

Rates of sea-level rise depend on several factors, including the rate of glacier and ice cap melting, the warming of ocean waters, and isostatic rebound, which is the vertical movement of the Earth's crust. <sup>110</sup> A national overview of coastal sensitivity to sea-level rise and associated storm impacts demonstrated that the Atlantic region has some of Canada's most severely threatened coastal areas. <sup>111</sup> Approximately 80% of the coastline is considered highly sensitive. Its most sensitive coastlines are generally low-lying areas with salt marshes, barrier beaches, and lagoons.

Over the past century, sea level in the Atlantic region has been rising; several harbours have experienced average rise rates of between 22 and 32 cm/century (Figure 22). Average sea level along the coastline of eastern Quebec rose by 17 cm over the last century. 107, 112 A portion of sealevel rise is likely due to land subsidence after glacier retreat, but much is due to sea-level rise from changing climate. For example, from 1911 to 2005, the annual mean sea level at Charlottetown rose at an average rate of 32 cm/century. 107 Of this, approximately 20 cm/century was likely due to land subsidence after glacier retreat and the remaining 12 cm/century was due to sea-level rise. 107, 113

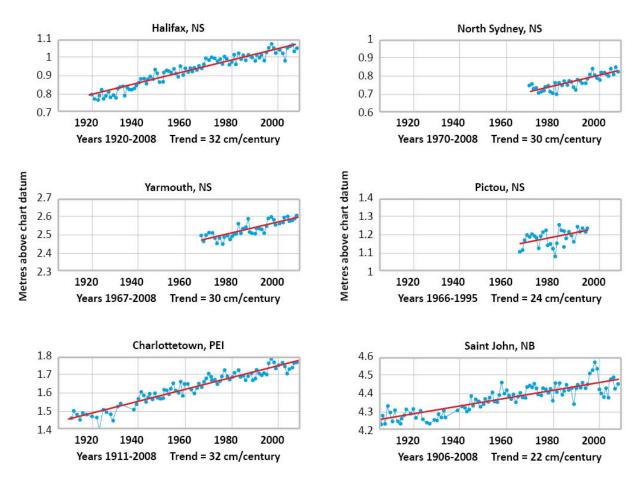


Figure 22. Trend in annual mean water level in six harbours in the Atlantic Maritime Ecozone $^{+}$ . Source: CBCL Limited,  $2009^{76}$  using data from Marine Environmental Data Service, Ottawa,  $2008^{114}$ 

One of the primary impacts of rising sea levels is an increase in coastal retreat or coastal erosion. Although coastal erosion is a natural phenomenon, rising sea levels as well as other climate change-related impacts to physical and climatic processes will accelerate erosion rates in parts of the AME, such as the Gulf of St. Lawrence. Accelerated coastal erosion is correlated with changes in climatic variables such as increased storm frequency, shorter ice season, more freeze/thaw cycles and winter rain events, and higher sea levels. The most sensitive areas to coastal erosion within the AME are on the Gaspé Peninsula, at the entrance of the Baie des Chaleurs, and around PEI and Îles de la Madeleine.

In some areas of PEI, there is already evidence of a significant increase in coastal erosion rates. For example, erosion rates at Pigots Point, Savage Harbour were 1.4 m/yr from 1968 to 1981 and 3.2 m/yr from 1981 to 1990. This is not necessarily the case throughout the AME, however. In 2006, Environment Canada quantified sea-level rise, storm surge, and coastal erosion on the region's Gulf of St. Lawrence coastal zone and found that coastal retreat rates for southeastern New Brunswick did not increase significantly during the second half of the 20th century. 107

In addition to erosion, other impacts on ecosystems from sea-level rise include higher and more frequent flooding of wetlands and adjacent shores, expanded flooding during severe storms and high tides, accelerated coastal (dune and cliff) retreat or erosion, breaching of coastal barriers and destabilization of inlets, saline intrusion into coastal freshwater aquifers, and damage to coastal infrastructure. Increased storm surge activity also has implications for coastal erosion and flooding (see Natural disturbance section on page 73).

Key finding 7 Theme Biomes

# Ice across biomes

### **National key finding**

Declining extent and thickness of sea ice, warming and thawing of permafrost, accelerating loss of glacier mass, and shortening of lake-ice seasons are detected across Canada's biomes. Impacts, apparent now in some areas and likely to spread, include effects on species and food webs.

Although ice is not a defining feature of the AME, it can provide important habitat for species adapted to living in, under, and on top of ice, and provide crossing points for land animals, and help to regulate water circulation. The timing and duration of ice cover on rivers, lakes, and the ocean are important factors in the types of plants and animals that water bodies can support.

#### River and lake ice

Information on overall trends in river and lake ice break-up and freeze-up in the AME was limited and inconclusive, <sup>63, 119-121</sup> and trends were limited to individual rivers or lakes (Table 7). <sup>122</sup> Of the ten sites covered by a recent analysis of data from the volunteer IceWatch program, only one trend, toward a later ice thaw date, was detected from 1950 to 2005. <sup>122</sup>

Table 7. Trends in lake freeze-up and break-up dates from studies in the Atlantic Maritime Ecozone +.

Freeze up	Dates	Change over time period	Trend per year	Significance
Grand Lake, NB <sup>120</sup>	1952–1980	17.4 days earlier	0.6 days/year	<0.1
Lake Utopia, NB <sup>120</sup>	1971–2000	37.5 days later	1.25 days/year	<0.001
Break up				
Lake Utopia, NB <sup>120</sup>	1961–1990	15.6 days earlier	0.5 days/year	<0.01
Miramichi River, NB <sup>121</sup>	1829–1955	7.3 days earlier/100 years		<0.01
Saint John River, NB <sup>123</sup>	1950s-1980s	15 days earlier		?

Sources are indicated as reference numbers after the name of the lake/river.

Prowse and Culp¹²⁴ provided a review of the effects of ice on instream ecological communities. In general, the life cycles of many aquatic organisms are both directly and indirectly influenced by ice through factors such as ice cover duration, instream temperatures, and hydrological variability. For example, Cunjak *et al.*¹²⁵ demonstrated that the interannual variability in the juvenile survival of Atlantic salmon (*Salmo salar*) in Catamaran Brook, NB, generally improved with increasing average winter flow but the lowest measured survival was associated with an atypical winter breakup and ice jam triggered by rain-on-snow snowmelt events.

### Sea ice

Sea ice is important in the AME as it is believed to have a dampening effect on wave action that causes coastal erosion and flooding. <sup>107</sup> In parts of the AME that have sea ice annually, ice cover varies from year-to-year; cycles are apparent and have some correlation with the North Atlantic Oscillation, a phenomenon of fluctuations in the difference in atmospheric pressure between the Icelandic Low and the Azores High, which in turn influences wind strength and direction. In the Gulf of the St. Lawrence, sea ice has shown a tendency toward decreasing ice cover and length of the ice season but these trends were not significant (Figure 23). <sup>107</sup> Saucier and Senneville<sup>126</sup> suggest that winter sea ice will be gone from the Gulf of St. Lawrence before the end of the 21<sup>st</sup> century, which could result in significant coastal erosion, including loss of coastal marshes (see Coastal section on page 29).

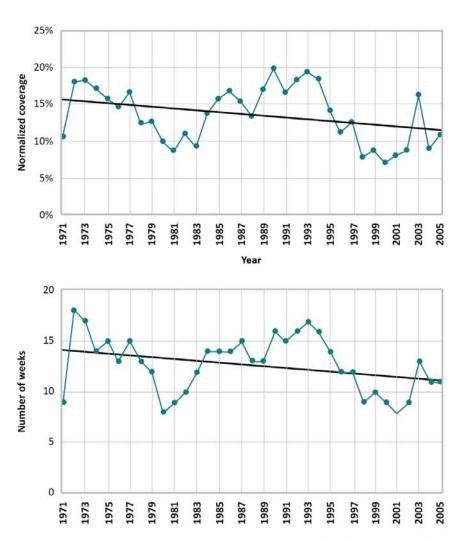


Figure 23. Trend in total accumulated ice coverage (top) and length of season (cover >10%) (bottom) for the Gulf of St. Lawrence, 1971–2005.

Source: Forbes et al., 2006<sup>107</sup>

# THEME: HUMAN/ECOSYSTEM INTERACTIONS

**Key finding 8** 

**Theme** Human/ecosystem interactions

## **Protected areas**

#### **National key finding**

Both the extent and representativeness of the protected areas network have increased in recent years. In many places, the area protected is well above the United Nations 10% target. It is below the target in highly developed areas and the oceans.

In 2009, the protected areas in the AME consisted of 617 protected areas that covered 5.3% of the landbase (Figure 24). This total was comprised of 438 protected areas in IUCN categories I–IV (10,963 km²; 4.9% of the AME), 172 protected areas in IUCN categories V–VI (796 km²; 0.4% of the AME), and 7 protected areas (<0.01% of the AME) not classified by IUCN category (Figure 25). IUCN categories I–IV include nature reserves, wilderness areas, and other parks and reserves managed for conservation of ecosystems and natural and cultural features, as well as those managed mainly for habitat and wildlife conservation. IUCN categories V–VI focus on sustainable use by established cultural tradition within the protected area. In 1992 (the signing of the Convention on Biological Diversity), 1.6% of the AME was protected.

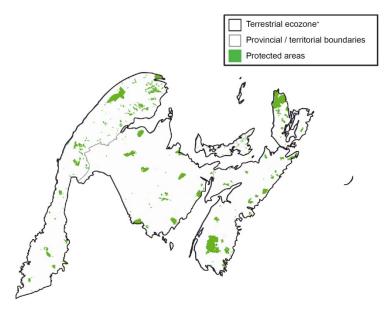


Figure 24. Distribution of protected areas in the Atlantic Maritime Ecozone<sup>+</sup>, May 2009. Source: Environment Canada, 2009;<sup>128</sup> using Conservation Areas Reporting and Tracking System (CARTS) data (v.2009.05)<sup>129</sup> provided by federal, provincial, and territorial jurisdictions

<sup>iv</sup> There are 2,100 km<sup>2</sup> of protected land in the AME with no information on the year established. Even if protected prior to 1992, the percentage of the AME protected prior to 1992 would still be 1.6%.

Prior to 1936, there was only 4 km² in category IV consisting of a single site, Amherst Point Migratory Bird Sanctuary in Nova Scotia, established in 1927. The total amount of protected area increased from under 1,000 km² in 1936, to just over 3,000 km² in 1992, and to over 11,000 km² in 2009 (Figure 25). The creation of seven national parks in the AME was responsible for most of the increases from 1936 to the 1980s. Cape Breton Highlands National Park in northern Nova Scotia, the first and largest national park in the AME (949 km²), was established in 1936. Kejimkujik National Park and National Historic Site in southern Nova Scotia, the second largest protected area in the region (404 km²), was opened in 1974. Recent additions since 1992 have been predominantly provincial parks and protected areas, mainly in Quebec and Nova Scotia.

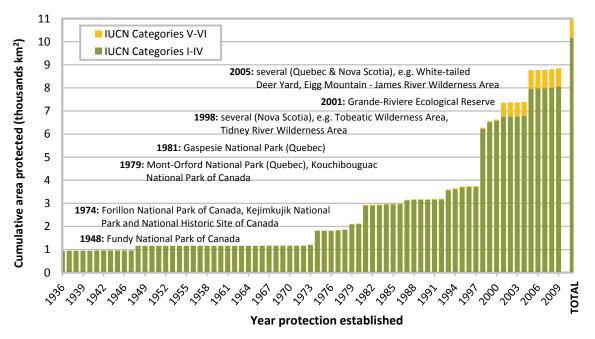


Figure 25. Growth of protected areas in the Atlantic Maritime Ecozone<sup>+</sup>, 1936–2009. Data provided by federal and provincial jurisdictions, updated to May 2009. Only legally protected areas are included. IUCN (International Union for Conservation of Nature) categories of protected areas are based on primary management objectives (see text for more information). The last bar marked 'TOTAL' includes protected areas for which the year established was not provided.

Source: Environment Canada, 2009;<sup>128</sup> using Conservation Areas Reporting and Tracking System (CARTS) data (v.2009.05)<sup>129</sup> provided by federal, provincial, and territorial jurisdictions

# **Invasive non-native species**

#### **National key finding**

Invasive non-native species are a significant stressor on ecosystem functions, processes, and structure in terrestrial, freshwater, and marine environments. This impact is increasing as numbers of invasive non-native species continue to rise and their distributions continue to expand.

Non-native species are plants, animals, or other organisms introduced by human activity into areas outside of their natural ranges. Non-native species are considered *invasive* when their introduction or spread threatens native species or ecosystems, or has the potential to cause considerable harm to the economy or society (e.g., due to their impacts on agricultural crops or forestry). Invasive non-native species are recognized as one of most significant threats to native biodiversity. Since the AME borders the ocean and has many ports, it has often been a point of entry for invasive non-native species. Comprehensive trend data do not exist for the AME, so this section presents a few examples where data exist.

# Invasive plants

The floras of Nova Scotia, New Brunswick, and PEI are composed of 37, 34, and 35% non-native species, respectively (Figure 26). However, there are currently only a few non-native plant species in the AME that appear to be having widespread negative impacts on native biological diversity. Only 36% of reported non-native species in the AME (not including Quebec) are known to be widely established (Figure 27). In general, the AME was less affected by invasive non-natives than the Great Lakes region or the heavily settled parts of the northeast United States. 66, 131

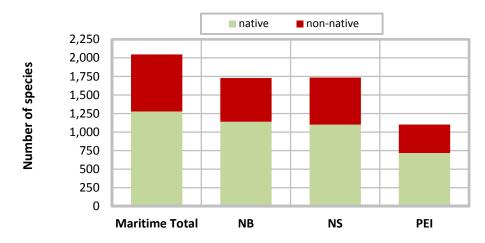


Figure 26. Total number of native and non-native plant species in the Maritime provinces, 2001. Data from Quebec is not included.

Source: adapted from Atlantic Canada Conservation Data Centre, unpublished data<sup>133</sup>

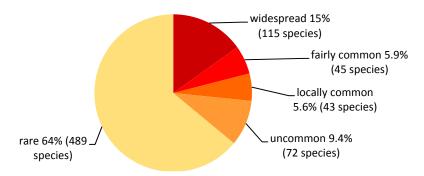


Figure 27. Abundance of non-native plant species in the Maritime Provinces, 2001. Data from Quebec was not included.

Source: adapted from Blaney, unpublished data<sup>132</sup>

Two non-native species in particular represent serious and broad threats: European common reed (*Phragmites australis* ssp. *australis*) and glossy buckthorn (*Frangula alnus*, also known by the synonym *Rhamnus frangula*). Other species of concern are Oriental bittersweet (*Celastrus orbiculatus*), purple loosestrife (*Lythrum salicaria*), Japanese knotweed (*Polygonum cuspidatum*), and garlic mustard (*Alliaria petiolata*). Another serious issue in the AME is the invasion of reed canary grass (*Phalaris arundinacea*) in streambeds and river shores.

# Invasive non-native insects and diseases

Non-native insects and diseases have had significant ecological impacts,<sup>134</sup> especially on forest ecosystems.<sup>10</sup> Trend data do not exist but important diseases include white pine blister rust, beech bark disease, and Dutch elm disease. There are 12 major introduced insect pest species in Nova Scotia with introduction dates ranging from the 1890s to 2000 (Table 8). Most of them arrived along the Eastern Seaboard in shipments from Europe over the last century. Many of these affect the entire AME.<sup>134</sup> Two examples are highlighted below.

Table 8. Major invasive non-native insects, and diseases in Nova Scotia, including year of introduction, location of first introduction to North America, and preferred host species, 1890s–2000.

Insect/disease	Year	Location of first introduction to North America	Preferred host
Beech bark disease	1890s	Halifax, NS	American beech
			(Fagus grandifolia)
Balsam woolly adelgid	1910s	Western Nova Scotia	Balsam fir
(Adelges piceae)			(Abies balsamea)
European spruce sawfly	1922	Ottawa, ON	Spruce
(Gilpinia hercyniae)			( <i>Picea</i> spp.)
Mountain-ash sawfly	1926	New York	Mountain ash
(Pristiphora geniculata)			(Sorbus americana)
White pine blister rust	1929	Chester, NS	Eastern white pine
(Cronartium ribicola)			(Pinus strobus)
European winter moth	1950	Nova Scotia	Oak
(Operophtera brumata)			(Quercus spp.)
Dutch elm disease	1969	Liverpool, NS	American elm
			(Ulmus americana)
Gypsy moth	1981	Yarmouth, NS	Hardwoods
(Lymantria dispar)			
Spruce longhorn beetle	2000	Halifax, NS	Red spruce
(Tetropium fuscum)	124		(Picea rubens)

Source: adapted from Neily et al., 2007<sup>134</sup>

#### Beech bark disease

Beech bark canker disease and its associated insect pathogen, beech scale (*Cryptoccoccus fagisuga*), have effectively eliminated large American beech trees from tolerant hardwood forests of PEI, Nova Scotia, and southern New Brunswick.<sup>135</sup> Beech was once a major component of these forests. Both the insect and the disease it carried were introduced from Europe through the Port of Halifax and were established in New Brunswick by 1927.<sup>136</sup> Beech trees that are genetically resistant to infection survive in infected areas.<sup>136</sup> Considering that beech was one of the most common species in the region, the disease has altered Acadian forest composition and has affected the availability of mast (or beechnuts) which is harvested as a food source.<sup>16</sup>

#### Brown spruce longhorn beetle

In contrast to the spruce budworm (described in the Natural disturbance section on page 76), the brown spruce longhorn beetle (*Tetropium fuscum*) is a new non-native invasive forest pest. It has been present since 1990 in Point Pleasant Park in Halifax, <sup>10, 137-139</sup> and remains localized to that area. <sup>138, 139</sup> The potential impact of the species on the forests of the AME and the rest of Canada is uncertain. Though the beetle has infested mainly red spruce in Point Pleasant Park, it is capable of attacking all spruce species native to Canada, other softwood species such as firs, pines, and larches, and occasionally hardwood species. <sup>10</sup>

# Invasive non-native freshwater species

Invasive non-native freshwater species can affect biodiversity and the health of aquatic ecosystems through competition with native aquatic species. .

#### **Smallmouth bass**

Smallmouth bass (*Micropterus dolomieu*) was originally found in lakes and rivers of eastern and central North America. As a result of widespread introductions, it is now found in south and central New Brunswick and Nova Scotia and east from southern Manitoba to Quebec.<sup>140</sup> It moved into New Brunswick in the 1870s<sup>141</sup> and, between 1905 and 1948, was stocked in six lakes in the south. As of 2009, it was found in over 70 lakes and 31 rivers in New Brunswick due to unauthorized stocking and natural spread.<sup>142</sup> In 2008, it was first recorded in the Miramichi River drainage, a world-class Atlantic salmon river.<sup>142</sup> In Nova Scotia, smallmouth bass was introduced into 11 lakes between 1942 and 1953 through stocking<sup>143</sup> and again between 1967 and 1984 (Figure 28).<sup>144</sup> The distribution today includes most of the south and central portion of the province.<sup>144</sup>

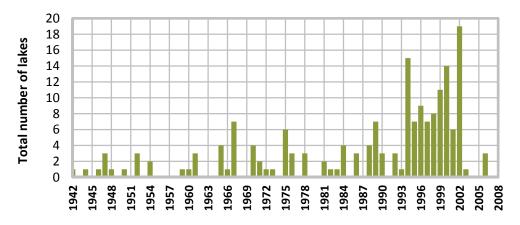


Figure 28. Number of lakes with first known occurrences of smallmouth bass in Nova Scotia, 1942–2008. Source: adapted from LeBlanc, 2009<sup>144</sup>

In the AME, smallmouth bass are an effective predator and competitor of other fish, including native Atlantic salmon.<sup>145</sup> The establishment of smallmouth bass in new systems has been shown to alter food webs and resulted in changes in species composition, relative abundance, and habitat use of fish assemblages, particularly for small-bodied fish species.<sup>143, 146</sup>

## Didymo

Didymo (*Didymosphenia geminata*) is a single-celled, microscopic freshwater alga endemic to rivers and lakes in boreal and mountainous regions of the Northern Hemisphere. When the algae produces profuse amounts of stalks, nuisance blooms can develop.<sup>147</sup> Since its first observation in the Matapedia River in the summer of 2006, it was observed in several rivers in Bas-Saint-Laurent, Gaspé Peninsula, and northern New Brunswick.<sup>148</sup> Didymo increased benthic

macroinvertebrate densities thus affecting the aquatic food web of the Matapedia River from 2006 to 2007.<sup>149</sup> When a bloom occurs, the mat can grow to cover extensive areas of stream bed and exposed substrate, causing significant harm to ecosystems.<sup>150</sup> The full extent of impacts on the ecosystem, including salmon, is still uncertain<sup>147, 148</sup>

# Invasive non-native marine species

## European green crab

Native to Europe and Northern Africa, the European green crab (*Carcinus maenus*) is one of the world's most successful invaders and has established on temperate coastlines on all continents. The main mechanism of spread has been through unintentional transport by the fishing vessel traffic and shipping, especially ships containing ballast water. The Green crabs are omnivores and feed voraciously on aquatic plants, bivalves, and particularly on molluscs, and are competitors for food with native predators and omnivores. In some parts of their introduced range, they have caused declines in other crab and bivalve species and are a threat to shellfish and fishing industries. In the Atlantic Maritime Ecozone, green crabs also threaten valuable eelgrass habitat; they can cut off eelgrass plants right at their shoots and are capable of affecting entire eelgrass meadows.

Other potentially important invasive non-native marine species include several tunicates, which are not discussed here.

## **Key finding 12**

Theme Human/ecosystem interactions

# **Nutrient loading and algal blooms**

#### **National key finding**

Inputs of nutrients to both freshwater and marine systems, particularly in urban and agriculture-dominated landscapes, have led to algal blooms that may be a nuisance and/or may be harmful. Nutrient inputs have been increasing in some places and decreasing in others.

Although the input of nutrients into groundwater and surface water occurs from many sources including urban, industrial, agricultural, and air-borne, this section focuses on risk of nitrogen and phosphorous loading from agricultural land. This section uses PEI as a case study of the trends in nitrates in groundwater and surface water and Quebec as a case study of trends in phosphorus in rivers.

# Nitrogen

Agricultural land in the AME have high levels of residual (or excess) soil nitrogen remaining on the land after inputs and outputs have been considered. Residual soil nitrogen levels

<sup>&</sup>lt;sup>ν</sup> Agricultural land in this indicator includes Cropland, Pasture, and Summerfallow categories from the Canadian Census of Agriculture.

increased on most cultivated land from 1981 to 2006 (Figure 29). In 2006, the AME had the second highest residual soil nitrogen values of all agricultural ecozones<sup>+</sup>, next to the Pacific Maritime Ecozone<sup>+</sup>. <sup>156</sup> As a result, the potential for leaching of nitrate out of soils and into water is high. In PEI, high nitrate concentrations in groundwater and surface water have become a serious issue for drinking water and ecosystem health. <sup>157</sup>

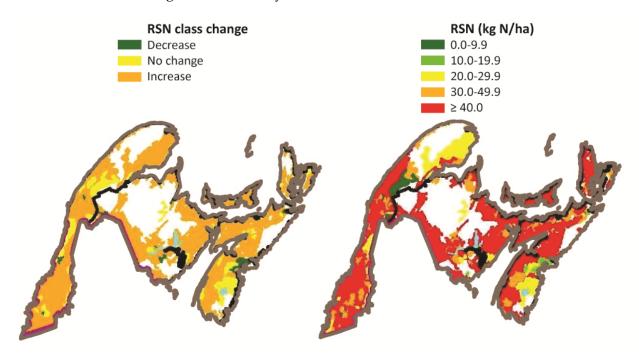


Figure 29. Change in Residual Soil Nitrogen (RSN) risk class from 1981 to 2006 (left) and risk classes in 2006 (right) for agricultural land in the Atlantic Maritime Ecozone<sup>†</sup>.

Agricultural land shown in this figure includes the Cropland, Improved Pasture, and Summerfallow categories from Canadian Census of Agriculture.

0.0-9.9 represents a very low risk class and >= 40 represents a very high risk class.

Source: Drury et al., 2011<sup>156</sup>

#### Nitrate levels in groundwater and surface water in PEI

Natural background nitrate levels are typically less than 2 mg/L. Aquatic biodiversity in rivers, streams, and estuaries is more sensitive to nitrate levels greater than 2-3 mg/L, which can inhibit growth, impair the immune system, and stress some species. Since the 1980s, PEI has experienced a steady increase in nitrate levels in groundwater. Average nitrate concentrations in groundwater from tested wells in PEI consistently exceeded 2 mg/L and remained above 3 mg/L between 1984 and 2007 (Figure 30). Nitrate concentrations in PEI's well water vary by watershed and patterns of contamination have remained consistent when compared between 2000–2005 and 2005–2008 (Figure 31). Generally, nitrate concentrations appear strongly associated with agricultural management practices in individual watersheds; watersheds with

the highest nitrate levels are in areas where the highest portion of the land is in potato production.

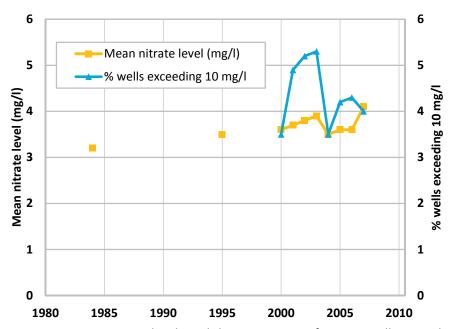


Figure 30. Mean nitrate levels and the percentage of private wells exceeding recommended nitrate concentrations, PEI, 1984–2007.

There were no data from 1985–1994 and 1996–1999.

Source: PEI Department of Environment, Energy and Forestry, unpublished data<sup>158</sup>

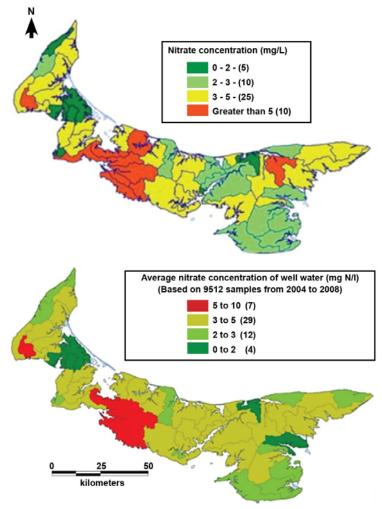


Figure 31. Change in average groundwater nitrate concentrations between 2000–2005 (top) and 2004–2008 (bottom) in watersheds in PEI.

Source: Commission on Nitrates in Groundwater, 2008 (2000–2005 data) $^{157}$ ; and Jiang, unpublished data (2004–2008 data) $^{159}$ 

Groundwater contributes as much as 65% of annual streamflow in a typical stream in PEI.<sup>160</sup> Nitrate enriched groundwater discharges to the local streams, leading to surface water contamination and aquatic ecosystem deterioration.<sup>161</sup> Monitoring data for all of PEI indicate nitrate concentrations of stream water have increased over time, and in some cases, have increased several-fold since the 1960s.<sup>161</sup> Excessive nutrient inputs can result in eutrophication, where macro algal overgrowth and dinoflagellate (phytoplankton) blooms deplete oxygen and/or release toxic substances, killing or choking out other wildlife. Algal overgrowth and dinoflagellate blooms can result from even relatively low levels of nitrate contamination (<2 mg/L), which lead to large-scale hypoxic or "dead zones".<sup>161, 162</sup> Between 2002 and 2008, 18 estuaries in PEI, the majority of which are on the north shore, were subject to recurring anoxic events (Figure 32).<sup>163</sup> Elevated nitrate in surface water has been suggested as one of the factors associated with the anoxia events.<sup>161</sup>

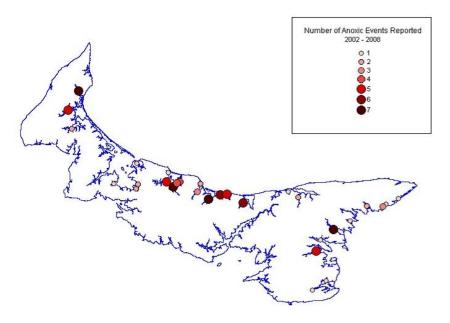


Figure 32. Number of anoxic events reported on Prince Edward Island between 2002 and 2008. Source: PEI Department of Environment, Fisheries and Forestry, unpublished data

# **Phosphorous**

According to monitoring data collected by Agriculture and Agri-Food Canada, from 1981–2006, risk of surface water contamination from soil phosphorus has increased in Canada, with an increasing percentage of agricultural watersheds at high and very high risk for contamination by phosphorus.<sup>164</sup> In Quebec and the Atlantic provinces, in particular, risk has gradually shifted from lower to higher risk classes since 1991 (Figure 33).<sup>164</sup> In terms of the amount of phosphorus in soils, the amount of farmland in Quebec and the Atlantic provinces exceeding the threshold value of 4 mg of phosphorus/kg of soil has increased from less than 2% in 1981 to over 33% in 2006.<sup>165</sup>

## Phosphorus concentrations in rivers in Quebec

In contrast to the results for agricultural lands above, phosphorus concentrations decreased by more than 50% at one station, between 0 and 50% at a second station, and were stable at three stations between 1988 and 1998 in rivers within the Quebec portion of the AME. 166 However, phosphorus levels also decreased at a series of control sites (witness stations) on rivers in the Appalachian Mountain lowlands when comparing 1979–2002 to 2000–2002. 166 These sites have watersheds with little to no human settlement. This suggests the factors influencing phosphorus concentrations in rivers may be declining naturally, regardless of human activities.

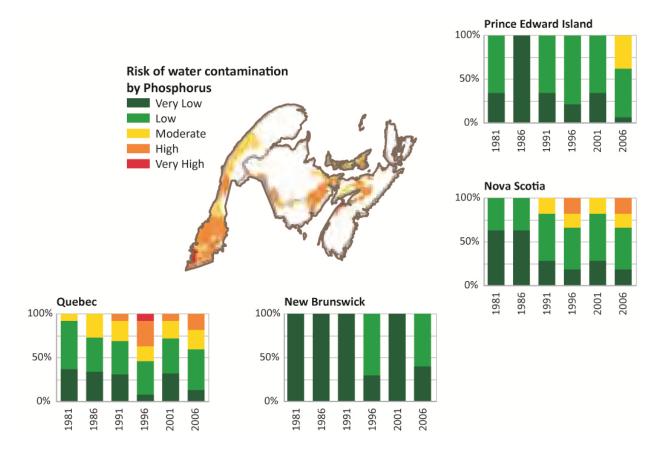


Figure 33. Risk of water contamination by phosphorous in agricultural watersheds under 2006 management practices in the Atlantic Maritime Ecozone $^{+}$  (map) and trend in the proportion of farmland in each risk class, 1981–2006, by province (bar graphs).

The Indicator of Risk of Water Contamination by Phosphorus (IROWC-P) was developed to assess the trends over time for the risk of surface water contamination by P from Canadian agricultural land at the watershed scale.

Quebec bar graph includes some area outside of the ecozone $^{+}$ .

Source: adapted from van Bochove et al., 2010<sup>164</sup>

## Algal blooms in Quebec

Blooms in blue-green algae (*Cyanobacteria*) have been linked to high phosphorus levels in surface water. <sup>167</sup> The number of lakes and rivers affected by blue-green algae in the Quebec portion of the AME has increased from three to 16 lakes between 2004 and 2008 (Figure 34).

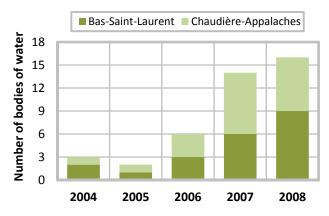


Figure 34. Number of lakes and rivers where blue-green algae was detected for Quebec administrative units that overlap with the Atlantic Maritime  $Ecozone^+$ , 2004–2008.

The Quebec administrative units that have the majority of their area in the AME are bas-Saint-Laurent and Chaudière-Appalaches.

Source: adapted from Ministère du Développement durable, de l'Environnement et des Parcs, 2009<sup>168</sup>

## **Key finding 13**

#### **Theme** Human/ecosystem interactions

# **Acid deposition**

### **National key finding**

Thresholds related to ecological impact of acid deposition, including acid rain, are exceeded in some areas, acidifying emissions are increasing in some areas, and biological recovery has not kept pace with emission reductions in other areas.

As a result of emission reductions, levels of sulphate and nitrate deposition in the AME decreased substantially between 1990 and 2004 (Figure 35).<sup>169</sup> Nonetheless, due to the poor buffering ability of its geology and soils, much of the AME is highly sensitive to acid<sup>170</sup> and atmospheric sulphur and nitrogen deposition exceeded critical loads in several areas from 1999 to 2003 (Figure 36).<sup>171</sup> Of particular concern is the potential long-term impact on forest health, for example, reduced growth rates, reduced productivity, increased mortality, and eventual changes in the composition of forest species.<sup>169, 171-173</sup>

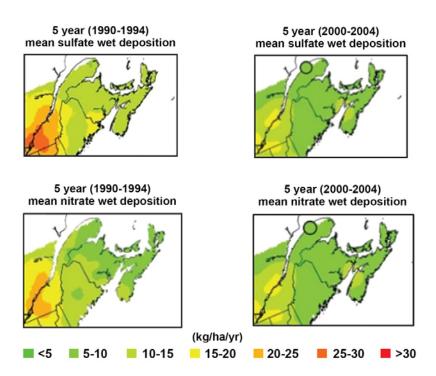


Figure 35. Trend in wet sulphate and wet nitrate deposition in the Atlantic Maritime Ecozone<sup>+</sup>, 1990− 1994 and 2000−2004.

Source: adapted from Commission for Environmental Cooperation, 2008<sup>169</sup>

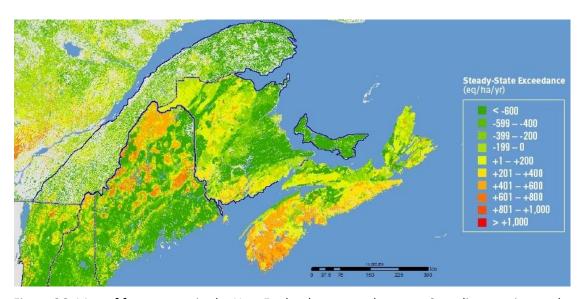


Figure 36. Map of forest areas in the New England states and eastern Canadian provinces where critical load has been exceeded due to acid deposition, ca. 1999–2003.

Data for atmospheric deposition rates from 1999–2003 in New England states and 1999–2002 in Quebec and the Atlantic provinces. Yellow, orange, and red areas are where sulphur and nitrogen have exceeded their critical loads. Green areas are where critical loads have not been exceeded.

Source: modified from New England Governors/Eastern Canadian Premiers Forest Mapping Group, 2007<sup>171</sup>

Another concern is the impact on fish and freshwater systems. The AME includes North America's most heavily affected region in terms of the percentage of fish habitat lost due to acid rain. <sup>170, 174</sup> Atlantic salmon are highly sensitive to acidity, and by 1996, 14 runs in coastal Nova Scotia were extinct because of water acidity, 20 were severely impacted, and a further 15 were lightly impacted. <sup>175</sup> There has been no measurable change in pH despite declines in sulphur dioxide emissions and recovery of water chemistry and ecology is expected to take several more decades in Nova Scotia than in other parts of Canada. <sup>170, 175, 176</sup> Recent research also suggests that the main driver of fish impacts is aluminum, which has been activated by acid deposition and reached levels that are toxic to fish. <sup>177</sup>

## **Key finding 14**

**Theme** Human/ecosystem interactions

# Climate change

#### **National key finding**

Rising temperatures across Canada, along with changes in other climatic variables over the past 50 years, have had both direct and indirect impacts on biodiversity in terrestrial, freshwater, and marine systems.

#### Trends in climatic variables

Table 9 summarizes significant trends in climatic variables in the AME from 1950 to 2007. The ecozone<sup>+</sup> is characterized by large variability at interannual and decadal to multi-decadal scales. Across the ecozone<sup>+</sup> as a whole, summer temperatures increased by only 1.1°C (Table 9, Figure 37). Relative to the rest of Canada, temperatures in the AME, Newfoundland Boreal, and Mixedwood Plains ecozones<sup>+</sup> rose the least over the 1950 to 2007 period.<sup>178</sup> In the AME, this was related to a widespread cooling trend over the northeast Atlantic ocean from approximately 1950 to 1980.<sup>179</sup>

Table 9. Summary of changes in climate variables in the Atlantic Maritime Ecozone<sup>+</sup>, 1950–2007.

Climate variable	Overall ecozone <sup>†</sup> trend (1950–2007)	Comments and regional variation
Temperature	<ul> <li>♠ of 1.1°C in summer</li> <li>No trend in spring, fall, or winter</li> </ul>	<ul> <li>Trends are consistent across ecozone<sup>+</sup></li> <li>Spring temperatures ↑ at two stations, near Sussex, NB, and Greenwood, NS</li> </ul>
Precipitation	<ul> <li>♠ of 18.6% in fall</li> <li>♠ number of days with precipitation in spring, summer, and fall</li> <li>No trend in ratio of snow to total precipitation</li> </ul>	<ul> <li>♠ in fall largely concentrated around northern portion of ecozone<sup>+</sup></li> </ul>
Snow	No trend in maximum snow depth or duration	<ul> <li>Snow cover season  by &gt;20 days at some stations (spring and fall)</li> <li>Maximum snow depth by &gt;40 cm at some stations</li> </ul>
Drought Severity Index	<ul><li>No trend</li><li>No extreme wet or severe drought years</li></ul>	<ul> <li>♠ of &gt;2 index units near Rimouski, QC (index ranges from 4 to -4)</li> <li>♦ of &gt;2 index levels near Saint John, NB</li> </ul>
Growing season	No change in length or start and end data	<ul> <li>♠ in length of growing season between 20–40 days and earlier start by 15–30 days at one station at the southern tip of Quebec</li> <li>Growing season started 0–15 days earlier at 3 stations around the Bay of Fundy</li> </ul>

Only significant trends (p<0.05) are shown.

Source: Zhang et al., 2011<sup>178</sup> and supplementary data provided by the authors

Fall precipitation increased as did the number of days with precipitation in spring, summer, and fall (Table 9), although there was some variation across stations (Figure 38). No overall trends in snow cover duration and annual maximum snow depth were found, however, trends were significant at a few individual stations where they consistently showed a shorter duration of snow cover (Figure 39) and lower maximum snow depths. Changes in precipitation have an impact on hydrology as discussed in the Lakes and rivers section on page 29.

Climate stations were well distributed across the AME and trends at individual stations were generally well reflected in the overall trends. There were some exceptions where individual stations showed significant changes that were different from the overall trends (see Table 9, Figure 37, Figure 38, and Figure 39).

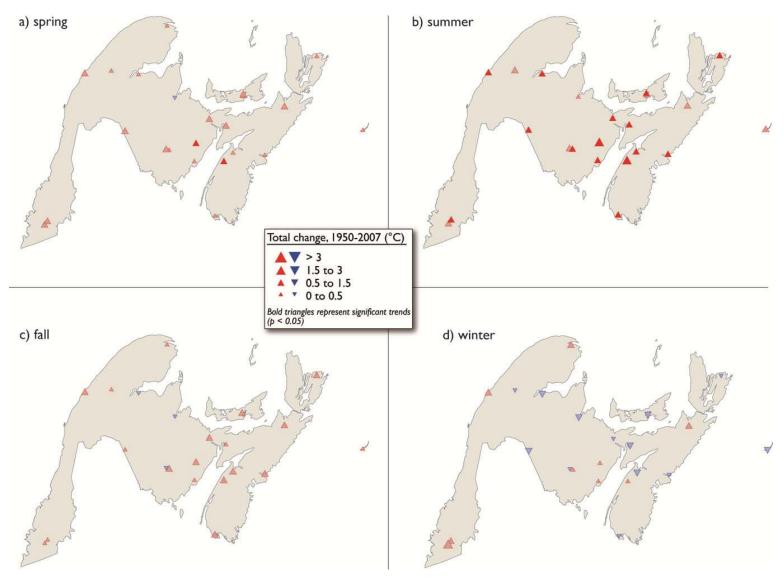


Figure 37. Change in mean temperatures in the Atlantic Maritime Ecozone<sup>+</sup>, 1950–2007, for: a) spring (March–May), b) summer (June–August, c) fall (September–November), and d) winter (December–February). Source: Zhang et al., 2011<sup>178</sup> and supplementary data provided by the authors

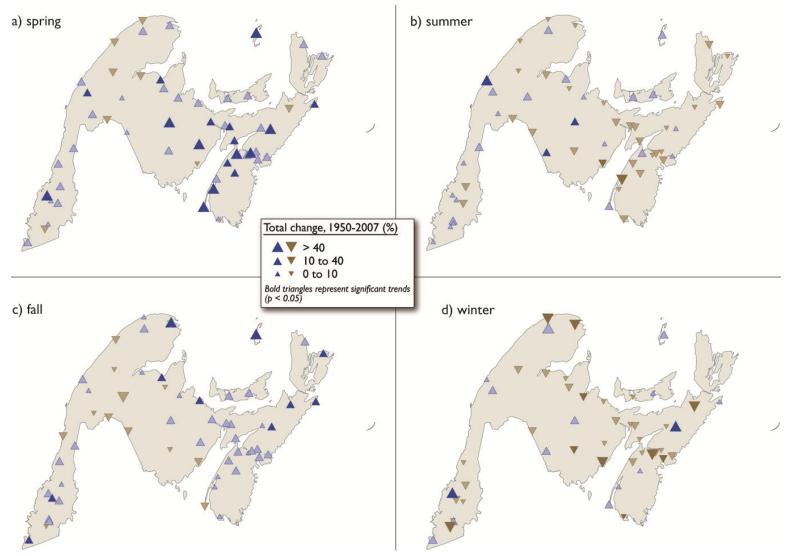


Figure 38. Change in the amounts of precipitation in the Atlantic Maritime Ecozone<sup>+</sup>, 1950–2007, for: a) spring (March–May), b) summer (June–August, c) fall (September–November), and d) winter (December–February).

Expressed as a percentage of the 1961–1990 mean.
Source: Zhang et al., 2011<sup>178</sup> and supplementary data provided by the authors

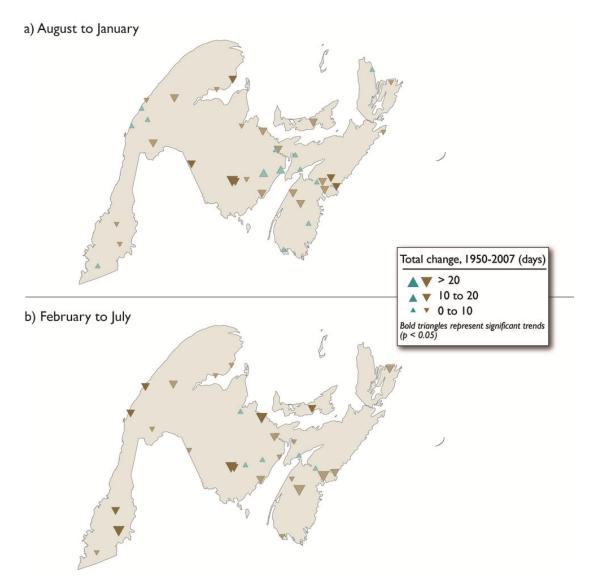


Figure 39. Change in snow durations (the number of days with ≥2 cm of snow on the ground) in the Atlantic Maritime Ecozone<sup>+</sup>, 1950–2007, in: a) the first half of the snow season (August−January), which indicates change in the start date of snow cover, and b) the second half of the snow season (February– July), which indicates changes in the end date of snow cover. Source: Zhang et al., 2011<sup>178</sup> and supplementary data provided by the authors

# **Future climate predictions**

Climate change is expected to have a range of effects of the AME. These include:

- Increased average annual air temperatures, although likely less than other parts of Canada;<sup>180</sup>
- Increased river water temperatures;<sup>181</sup>
- A longer, warmer growing season;<sup>181</sup>
- Decreased sea ice cover in the Gulf of St. Lawrence;<sup>126</sup>
- Changes in storm intensity and frequency;<sup>182</sup> and Changes in forest composition (e.g., a reduction in the proportion of yellow birch and an expansion by white birch and poplar).<sup>26</sup>

Some fish species, such as Atlantic salmon, are cold-water species, and warmer waters could have a negative impact their growth. Warmer waters can increase salmon's susceptibility to disease and infection, increase mortality rates, and decrease the availability of suitable habitat. Modeling suggests that climate change could increase river water temperatures in the region by 2–5° C and produce more extreme low flow conditions. Research in the Miramichi River examined the relationship between climate, hydrological parameters, and the length of juvenile salmon (parr) and detected a significant decline in length. Fish length is an indicator of growth that also affects competition, predation, smoltification, and marine survival. This relationship was associated with the warming observed and the results suggest that future climate change will adversely affect juvenile salmon in the Miramichi River. 181

#### **Key finding 15**

**Theme** Human/ecosystem interactions

# **Ecosystem services**

### **National key finding**

Canada is well endowed with a natural environment that provides ecosystem services upon which our quality of life depends. In some areas where stressors have impaired ecosystem function, the cost of maintaining ecosystem services is high and deterioration in quantity, quality, and access to ecosystem services is evident.

Ecosystem services are the direct goods and indirect services from a healthy, natural environment that ensure human well-being. These include four different types of services: provisioning services, regulating services, supporting services, and cultural services. Provisioning services in the AME include forest products, water, food, and commercial freshwater fishing. Regulating services such as wastewater assimilation are important, as are the supporting services provided by wetlands. Ecosystems also contribute important cultural services, such as recreational fishing, hunting, outdoor recreation, and tourism.

Valuation of ecosystem goods and services accounts for ecosystem stocks and flows using biophysical or monetary measures. Basic economic analysis typically accounts for flows of goods from ecosystems including, for example, forest products, fish, food, and energy and

mineral resources. These are traded in economic markets and their value over time may serve as indicators of ecosystem status and trends. Other ecosystem goods and services, however, such as climate regulation, water purification, and waste assimilation are not traded in markets and are referred to as non-market goods and services.

The combined estimated value of ecosystem goods and services for the Atlantic provinces (excluding the Quebec portion of the AME, because it could not be easy separated out from other parts of Quebec) is over \$4.7 billion (Table 10).

Table 10. Summary of the estimated values of ecosystem goods and services in the Atlantic Maritime Ecozone $^{+}$ , excluding the Quebec portion.

Service	Year	Value (millions)**	Measure
Water	Various	\$2,434	Various
Forests	2006	\$466	GDP + farm value
Outdoor recreation	1996	\$463	Expenditures
Fishing (commercial)	2006	\$406	Landed value
Agriculture	2006	\$347	Added value
Tourism	2006	\$300	Expenditures
Wetlands*	2007	\$122	Choice experiment
Recreational fishing	2006	\$122	Expenditures
Total		\$4,753	

<sup>\*</sup> Wetland figures not included in total to avoid double counting

Source: Eaton, 2013<sup>4</sup> using data from various sources

# THEME: HABITAT, WILDLIFE, AND ECOSYSTEM PROCESSES

**Key finding 16** 

**Theme** Habitat, wildlife, and ecosystem processes

# Agricultural landscapes as habitat

### **National key finding**

The potential capacity of agricultural landscapes to support wildlife in Canada has declined over the past 20 years, largely due to the intensification of agriculture and the loss of natural and semi-natural land cover.

Although some biodiversity is lost when land is converted to agriculture, agricultural lands still contribute significant biodiversity values as the varied habitats on agricultural landscapes provide some or all of the requirements for many wildlife species.<sup>183</sup>

Agricultural lands<sup>vi</sup> comprised close to 10% of the AME in 2006 and were characterized generally by small-scale farming. With the exception of a few areas of higher production (e.g.,

<sup>\*\*</sup> Values converted to 2006 dollars

vi Agricultural land in this section was defined as cropland, summer fallow, pastureland, and woodlands and wetlands that were reported by farmers in the Canadian Census of Agriculture.

PEI, Annapolis–Minas Lowlands, Saint John River Valley), agricultural land made up a relatively small component of the broader landscape (Figure 40).<sup>184</sup> Most agricultural areas were made up of a diversity of cover types that included a considerable amount of natural and seminatural land. The relatively light agricultural footprint along with the presence of abundant, high-value habitat on agricultural land means that the influence of agriculture on habitat is much less here than in the major Canadian agricultural ecozones<sup>+</sup>.

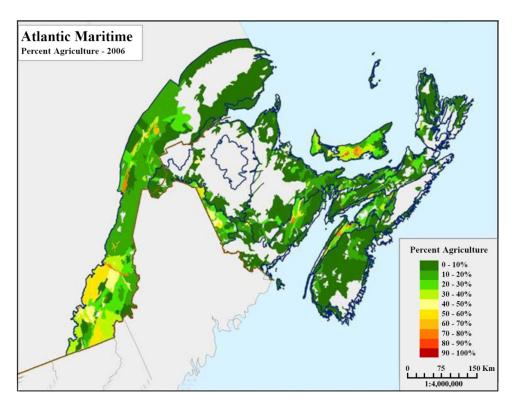


Figure 40. Percentage of land defined as agricultural in the Atlantic Maritime Ecozone<sup>+</sup>, 2006. Soil Landscapes of Canada polygons were the base unit used for this analysis. Source: Javorek and Grant, 2011<sup>184</sup>

From 1986 to 2006, the total agricultural land shrank by about 6% (from 22,000 to 20,800 km²). The share of "All Other Land" declined from approximately 49 to 47% of the total agricultural landscape. Tame Hay, the second most abundant cover type, expanded its share from 21 to 26%, while both Improved Pasture (9 to 5%) and Unimproved Pasture (9 to 6%) declined. The share of Other Crops expanded from 2 to 3%, mainly due to increased potato production on Prince Edward Island and in the Saint John River Valley (Figure 41).<sup>184</sup>

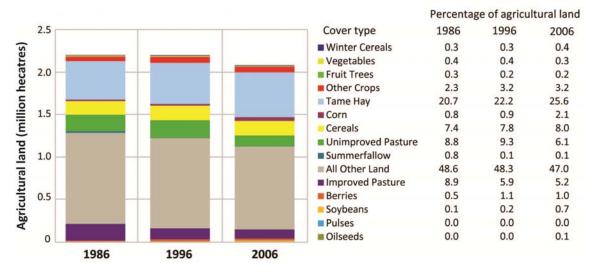


Figure 41. Total agriculture area, the amount of land per cover type (bar chart), and the relative percentage of each cover type (table) for the Atlantic Maritime Ecozone<sup>+</sup> in 1986, 1996, and 2006. Source: Javorek and Grant, 2011<sup>184</sup>

# Wildlife habitat capacity on agricultural land

A total of 292 species (215 birds, 52 mammals, 9 reptiles, and 16 amphibians) potentially use this agricultural landscape, with 88% associated with wetland, riparian, shelterbelts, woodland, old field, and idle land (All Other Land category). The All Other Land category was the dominant land cover type making up close to half of the total agricultural land base. The capacity of agricultural landscapes to provide habitat for wildlife was calculated for the years 1986, 2001, and 2006 using a model that ranked land cover types based on potential uses (e.g., breeding and reproduction, migration, wintering) and value (primary, secondary, or tertiary) for different species into ten categories (see legend in Figure 42). <sup>184</sup> In 2006, average wildlife habitat capacity on agricultural land was rated as high despite a significant decline since 1986 (Figure 42). Between 1986 and 2006, habitat capacity decreased on 43% of agricultural land, increased on 28%, and was constant on 29% (Figure 43). Declining habitat capacity trends were associated with a number of areas reporting more intense agricultural activity. The significant decline resulted from a general expansion of the comparatively low habitat Cropland (32 to 36%) and a decline of cover types with higher value to wildlife. <sup>184</sup> Despite this decline, average wildlife habitat capacity in the AME remained high.

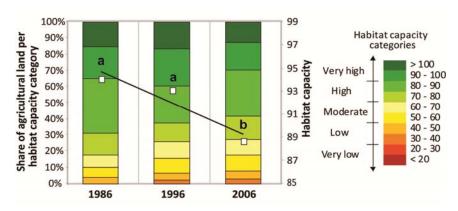


Figure 42. The share of agricultural land in each habitat capacity category (left axis, stacked bars) and the average habitat capacity (right axis, points and line) for the Atlantic Maritime Ecozone<sup>†</sup> in 1986, 1996, and 2006.

Years with different letters indicate a statistically significant difference.

Source: Javorek and Grant, 2011<sup>184</sup>

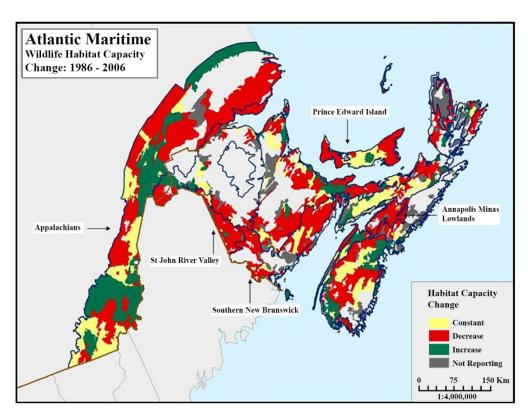


Figure 43. Change in wildlife habitat capacity on agricultural lands in the Atlantic Maritime Ecozone<sup>+</sup>, 1986–2006.

All Soil Landscapes of Canada (SLC) polygons with >5% agricultural land were included in the analysis. Source: Javorek and Grant, 2011<sup>184</sup>

# Soil erosion on cropland

Occupying only 4% of the total land area, cropland<sup>vii</sup> in the AME has some of the highest erosion risk on agricultural land in Canada due to intensive tillage and a climate that poses a high threat of water erosion of unprotected soils in some areas.<sup>185</sup> However, the risk of soil erosion declined in the AME from 1981 to 2006. McConkey *et al.*<sup>185</sup> found that 36% of the cropland was classified as having unsustainable erosion risk in 2006 (Figure 44), down from 41% in 1981. In 2006, 18% of agricultural land was at moderate to very high erosion risk compared to 20% in 1981.

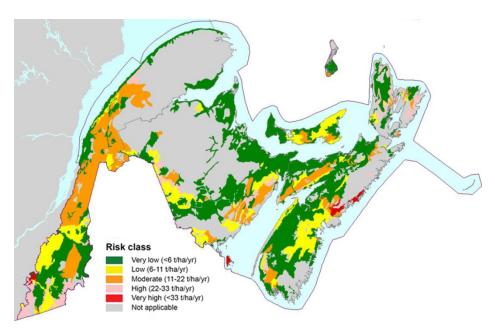


Figure 44. Soil erosion risk classes for cropland in the Atlantic Maritime Ecozone<sup>+</sup>, 2006. All Soil Landscape of Canada polygons containing >5% cropland were included in the analysis and entire polygons are shown on the map.

Source: McConkey et al., 2011<sup>185</sup>

# Birds of grassland and other open habitats

Grassland birds, which include birds of some agricultural habitats such as hayfield, pastures and rangeland, and birds of other open habitats, which include agricultural lands not included in the grassland category and abandoned fields, have declined significantly (Figure 45). Vesper sparrow (*Pooecetes gramineus*), bobolink (*Dolichonyx oryzivorus*), and eastern meadowlark (*Sturnella magna*) populations declined by over 75% since the 1970s. Many aerial-foraging insectivores, included in the other open habitat category, declined as a group.<sup>42</sup>

vii Cropland in this analysis also includes areas defined as Improved Pasture and Summerfallow in the Census of Agriculture. See McConkey *et al.*, 2011<sup>185</sup> for more information.

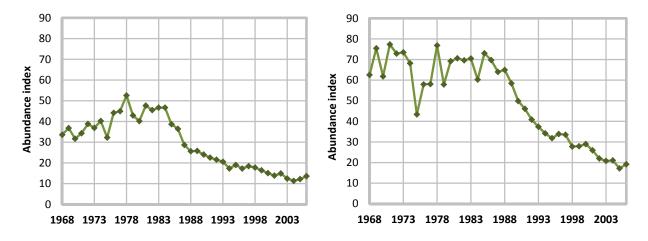


Figure 45. Annual indices of population change in bird assemblages for grassland habitat (left) and other open habitats (right) in the Atlantic Maritime Ecozone $^+$ , 1968–2006.

Grassland habitats include native grasslands and some agricultural habitat such as hayfield, pastures and rangeland. Other open habitats include open country, including species of agricultural landscapes not considered in grassland. The index is an estimate of the average number of individual birds that would be counted on a randomly selected route by an average observer in a given year.

Source: Downes et al., 2011<sup>42</sup>using data from the Breeding Bird Survey<sup>43</sup>

#### **Key finding 17**

Theme Habitat, wildlife, and ecosystem processes

# Species of special economic, cultural, or ecological interest

## **National key finding**

Many species of amphibians, fish, birds, and large mammals are of special economic, cultural, or ecological interest to Canadians. Some of these are declining in number and distribution, some are stable, and others are healthy or recovering.

Species of particular interest within the AME include the Atlantic (Gaspésie) population of woodland caribou and Atlantic salmon. Some landbirds that use the AME have declined , in some cases due to pressures elsewhere in their migratory ranges.

In the past 150 to 200 years, some of the largest mammals were extirpated from the AME. including wolf (*Canis lupus*), cougar (*Felis concolor*), and wolverine (*Gulo gulo*). Woodland caribou (*Rangifer tarandus caribou*) has been reduced to a single endangered population. The wolf's ecological niche has largely been filled by the coyote (*Canis latrans*), and that of the caribou has been filled to some degree by the white-tailed deer (*Odocoileus virginianus*). Beaver (*Castor canadensis*) were nearly extirpated 200 years ago due to overharvest, but have since recovered.

## Woodland caribou

The Atlantic-Gaspésie population of the woodland caribou is an isolated relict population that formerly ranged more broadly in the AME. Prior to European settlement, woodland caribou were commonly found throughout much of Nova Scotia and New Brunswick and were present on PEI. Extirpation from these three provinces was well underway by the 1830s. Caribou were extirpated from Nova Scotia by 1912, 187 New Brunswick by the 1930s, 16 and from PEI much earlier. Efforts to re-establish caribou on their historic ranges in Nova Scotia failed because of fatal infections with *Parelaphostrongylus tenuis*, a brain worm carried by the more recently established white-tailed deer. 188

The current population is found only in and adjacent to Gaspésie National Park of Quebec.<sup>189</sup> It is at risk from predation and habitat loss, and its low numbers and restricted range make it susceptible to chance catastrophic events.<sup>186</sup> Trend data from 1983 to 2006 show an overall decline over this period, with a low population size of less than 100 individuals in 1999 (Figure 46).<sup>189</sup> In 2002, COSEWIC re-assessed the population and elevated its status from Threatened to Endangered; it is also listed on Schedule 1 of Canada's *Species at Risk Act*.

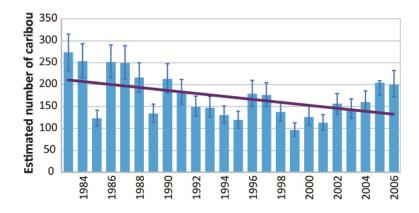


Figure 46. Trend in the estimated numbers of the Gaspésie woodland caribou population, 1983–2006. Source: Gaspésie Woodland Caribou Recovery Team, 2007<sup>189</sup> (http://www.mrnf.gouv.qc.ca/)

# Other ungulates

Other large herbivores in the AME include moose (*Alces alces*) and white-tailed deer. Nova Scotia mainland moose have declined by 20% to about 1,000 individuals since 1970 due to human intrusion into its habitat, hunting, climate change, and disease.<sup>190</sup> White-tailed deer are a recent arrival to the Maritimes and have been expanding. They have benefited from human modifications of the forested landscape, as well as extirpations or reductions of many of their predators.<sup>136</sup>

## Atlantic salmon

Atlantic salmon are broadly distributed in rivers throughout the AME. Populations are sensitive to a number of environmental factors including predation, fishing, and the availability of breeding habitat.<sup>191</sup> As was mentioned in the section (page 52), Atlantic salmon are also highly sensitive to acidity, and a high percentage of fish habitat has been lost in the region due to acid rain, with many runs in coastal Nova Scotia either extinct or heavily impacted.<sup>175</sup> Construction of dams has had an impact on salmon populations, and industrial and municipal effluents, as well as run-off from intensive agriculture, degrade water quality and reduce suitable breeding habitat for salmon. Invasive predators such as muskellunge (*Esox masquinongy*), smallmouth bass, and rainbow trout (*Oncorhynchus mykiss*) reduce juvenile salmon survival.

There is considerable variation in the status and trends in Atlantic salmon from one part of the AME to another. The Inner and Outer Bay of Fundy populations of Atlantic salmon were designated as Endangered by COSEWIC in 2001 and 2010, respectively. All survey data from the inner Bay of Fundy indicate that river-specific populations have suffered extreme declines since the 1970s and this population faces extinction. Estimates of declines are as high as 99% over 11 years (three generations) and greater than 99.6 % over 30 years. In 2003, fewer than 100 adults were estimated to have returned to the 32 rivers known to have contained salmon. Historically, as many as 40,000 salmon likely returned to these rivers. Although there is some uncertainty, it appears that offshore mortality of adult salmon is the primary threat to the Inner Bay of Fundy population.

Of 37 salmon rivers in the AME (18 in the Maritime provinces and 19 in Quebec) the five-year average population size increased in only three rivers, all on Cape Breton Island, from 1987 to 2005 (Figure 47). Abundance declined in all other rivers, with declines of over 95% in four rivers in the inner Bay of Fundy and a 99.8% decline in the St. Croix River in the outer Bay of Fundy. Trends in abundance vary throughout rivers in Quebec, though populations generally increased and declined in only two rivers.

The Miramichi River produces at least 20% of North American Atlantic salmon and more wild Atlantic salmon than any other North American river. The salmon in the Miramichi and Restigouche rivers are extremely important to overall Atlantic salmon populations because these two rivers contribute a disproportionate number of spawning fish to populations of maiden salmon that return to spawn in the rivers after spending two years at sea. Atlantic salmon abundance has declined in both rivers from 1987 to 2005 (Figure 47), although populations have shown some recovery since 2000.<sup>192</sup>

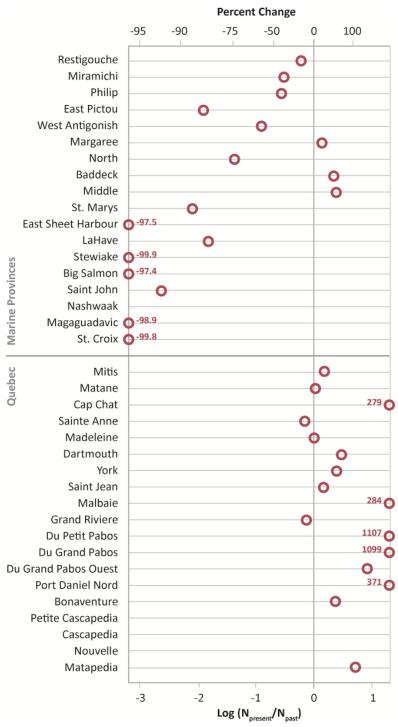


Figure 47. Changes in abundance of salmon populations for the Maritime provinces (top) and Quebec (bottom), 1987–2005.

Scale on x axis is  $Log(N_{present}/N_{past})$ . Each point is the change in five-year average population size. Points outside the graph's range are labelled with their value.

Source: modified from Gibson et al., 2006<sup>192</sup>

Numbers of Atlantic salmon in rivers in PEI also declined. The fish were thought to occur in about 70 rivers in PEI prior to European settlement. By 1960, this had declined to approximately 55 rivers, and a comprehensive study in 2000–2002 found salmon in just 33 rivers. In 2008, 11 more rivers no longer had salmon and populations in 7 others were very low.<sup>194</sup>

#### American eel

The American eel (Anguilla rostrata) is an example of a once abundant species that is now listed as Threatened by the Committee on the Status of Endangered Wildlife in Canada (COSEWIC). Since the 1970s, populations have declined by 99% in the upper St. Lawrence, 195 and less extreme declines have been observed in both the lower St. Lawrence and Gulf of St. Lawrence. 196, 197 The long life span of American eels, combined with their vast distances of up to 4,500 km, make them vulnerable to a wide range of stressors, such as mortality in hydroelectric turbines, physical barriers such as dams, overharvesting, and habitat alteration. Climate change, resulting in changes to ocean currents that carry eel larvae from the spawning grounds, may also contribute to population declines. American eels once provided both subsistence and commercial fisheries in Canada. 197

In the Atlantic Maritime Ecozone+, trends in American eel populations have been mixed. Electrofishing surveys have been conducted regularly in six major rivers with available time series of data ranging from 15 to 45 years. While four rivers in New Brunswick (Miramichi, Restigouche, Nashwaak, and Big Salmon rivers) saw above average abundance in the 2000s, abundance in two rivers in Nova Scotia has strongly declined, by about 75% in the St. Marys River from 1998 to 2009 and by 86% in the LaHave River from 2000 to 2009.

# Freshwater fish

Between 1979 and 2008, the number of freshwater and diadromous fish taxa classified as imperilled in the AME by the American Fisheries Society tripled from three to nine species (Figure 48, Table 11). Rainbow smelt was added as a result of improved status information while populations of striped bass and Atlantic salmon were added due to the inclusion of discrete regional populations as of 2008.<sup>198</sup>

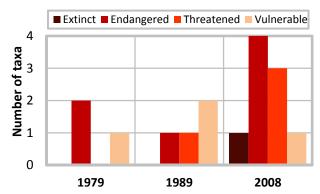


Figure 48. Trend in numbers of imperilled freshwater and diadromous fish taxa in each status category for North American ecoregions in the Atlantic Maritime Ecozone $^+$ , 1979, 1989, and 2008.

'Taxa' is used instead of 'species' because the list was updated to include discrete regional populations and infraspecific taxa. Previous lists may have underestimated the imperiled taxa because they did not include all designable units, only taxonomically recognized species.

Definitions of status categories differ slightly from COSEWIC and are described in Jelks et al. <sup>198</sup> Source: adapted from Jelks et al., 2008<sup>198</sup>

Table 11. Identification and status of imperilled freshwater and diadromous fish taxa in the Atlantic Maritime Ecozone $^{+}$ , 1979, 1989, and 2008.

English common name	Genus	Species	1979	1989	2008
Shortnose sturgeon	Acipenser	brevirostrum	Е	Т	Е
Atlantic sturgeon	Acipenser	oxyrinchus oxyrinchus	V	V	V
Striped bass (Bay of Fundy population)	Morone	saxatilis			Т
Striped bass (Southern Gulf of St. Lawrence population)	Morone	saxatilis			Т
Striped bass (St. Lawrence Estuary population)	Morone	saxatilis			Хр
Rainbow smelt (Lake Utopia, New Brunswick dwarf population)	Osmerus	mordax		V	Т
Atlantic whitefish	Coregonus	huntsmani	Е	Е	Е
Atlantic salmon (Bay of Fundy population)	Salmo	salar			E
Atlantic salmon (Gulf of Maine population)	Salmo	salar		100	Е

Xp = 'possibly extinct', E = 'endangered', T = 'threatened', V = 'vulnerable' as defined in Jelks et al.  $^{198}$  Source: adapted from Jelks et al.,  $2008^{198}$ 

#### Landbirds

All landbird species assemblages, except forest birds, declined from the 1970s to the 2000s, with the greatest declines in birds of grassland (includes species of agricultural habitats such as hayfields, pastures, and rangelands) and other open habitats (Table 12, see also Figure 45 in Agricultural landscapes as habitat section on page 60).

Table 12. Trends in abundance of landbirds for the Atlantic Maritime Ecozone<sup>+</sup>, 1970s to 2000s

Cuasias Assamblaga	Trend		BBS Abundance Index				
Species Assemblage	(%/yr)	Р	1970s	1980s	1990s	2000s	Change
Forest	-0.4%		221.6	218.3	208.1	187.1	-16%
Shrub/Successional	-0.6%	*	160.2	141.9	137.1	134.9	-16%
Grassland	-3.5%	*	39.9	38.2	19.5	13.3	-67%
Other Open	-3.5%	*	64.8	67.0	36.3	22.6	-65%
Urban / Suburban	-0.6%	*	179.7	162.0	157.3	154.9	-14%

P is the Statistical significance: \* indicates P < 0.05; n indicates 0.05 < P < 0.1; no value indicates not significant

Change" is the percent change in the average index of abundance between the first decade for which there are results (1970s) and the 2000s (2000-2006).

Source: Downes et al., 2011<sup>42</sup> using data from the Breeding Bird Survey<sup>43</sup>

## **Key finding 18**

Theme Habitat, wildlife, and ecosystem processes

# **Primary productivity**

#### **National key finding**

Primary productivity has increased on more than 20% of the vegetated land area of Canada over the past 20 years, as well as in some freshwater systems. The magnitude and timing of primary productivity are changing throughout the marine system.

The Normalized Difference Vegetation Index (NDVI), calculated from remote sensing data, is an indicator of the amount and vigour of green vegetation present on a landscape. Changes in NDVI are a proxy for changes in primary productivity. From 1985 to 2006, NDVI values increased for 33,408 km² (16.5%) and decreased for 720 km² (0.4%) of the AME.<sup>7</sup> The largest areas with increasing NDVI values were mixed forest along the Gaspé Peninsula and on Cape Breton Island (Figure 49).

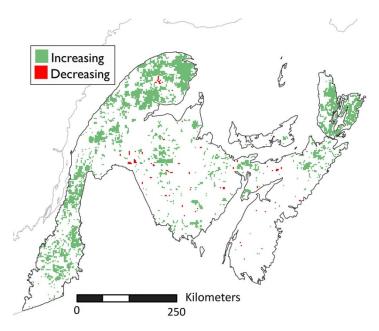


Figure 49. Change in the Normalized Difference Vegetation Index for the Atlantic Maritime Ecozone<sup>+</sup>, 1985–2006.

Trends are in annual peak NDVI, measured as the average of the three highest values from 10-day composite images taken during July and August of each year. Spatial resolution is 1 km, averaged to 3 km for analysis. Only points with statistically significant changes (p<0.05) are shown. Source: adapted from Pouliot et al, 2009<sup>199</sup> by Ahern et al., 2011<sup>7</sup>

Changes in NDVI can be attributed to climate change, land cover change, and land use or other management changes.<sup>7</sup> Increasing trends in parts of the AME may be associated with commercial logging that has increased the proportion of broadleaf trees, but more detailed studies would be needed to confirm this hypothesis. Because of the high proportion of deciduous and mixed deciduous forests in this ecozone<sup>+</sup>, NDVI values were in a higher range, close to the saturation point, making subtle changes difficult to detect.<sup>199-201</sup> In addition, the result of NDVI analyses in southeastern Canada (including the AME) are sensitive to the period being analyzed. Earlier time periods (such as 1982 to 1999) are more likely to show extensive increasing trends<sup>202-205</sup> while analyses of more recent periods (such as 1985 to 2006, as analyzed here) show less extensive positive trends or even some areas of negative trends.<sup>199, 206, 207</sup> More detailed land cover and vegetation productivity studies would be necessary to fully understand these trends.

## **Natural disturbance**

#### **National key finding**

The dynamics of natural disturbance regimes, such as fire and native insect outbreaks, are changing and this is reshaping the landscape. The direction and degree of change vary.

Natural disturbances include extreme weather events, fire, and insect outbreaks. Although fire was important within the AME historically, severe weather events and insect outbreaks are the dominant disturbance types today due, in part, to effective fire suppression. Spruce budworm is the most influential forest insect.

#### Extreme weather events

Since the AME borders the Atlantic Ocean, it is especially vulnerable to hurricanes and other tropical storms tracking up North America's Eastern Seaboard. The winds and tidal events associated with these storms can also lead to storm surges and flooding.

## **Tropical storms and hurricanes**

The frequency and severity of tropical storms and hurricanes has increased over the past three decades.<sup>208-211</sup> The average number of tropical cyclones—that is, hurricanes, tropical storms, and tropical depresssions—per year was 8.7 from 1900 to 1999, 9.9 from 1950 to 1999, and rose to 11.8 from 1991 to 2000, the highest 10-year average on record (Figure 50).<sup>211</sup> Other recent studies showed that the duration of tropical-storm events in the Atlantic region has increased by about 60% since 1949 and the annual peak-wind speed increased by about 50%.<sup>103</sup> Since 1975, the total dissipation of power (an index of a hurricane's potential destructiveness) has doubled.<sup>212</sup>

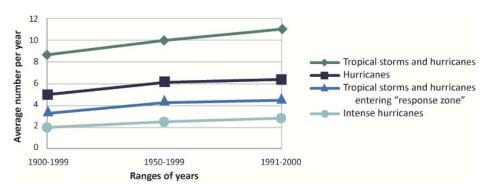


Figure 50. Trends in the average number of tropical cyclones in the Atlantic Basin, 1900–1999, 1950–1999, and 1991–2000.

Source: Environment Canada, 2002<sup>211</sup>

#### Storm surges and flooding

Storm surges and flooding are often associated with hurricanes and tropical storms and result from increased marine-wave action and heavy rainfall. They can have significant impacts on coastal ecosystems, including soil erosion and vegetation loss (see also Coastal section on page 29). Susceptibility to storm surges varies widely in the AME: some areas are likely to be more severely affected than others, depending on the nature of the coastline and degree of exposure (Figure 51).

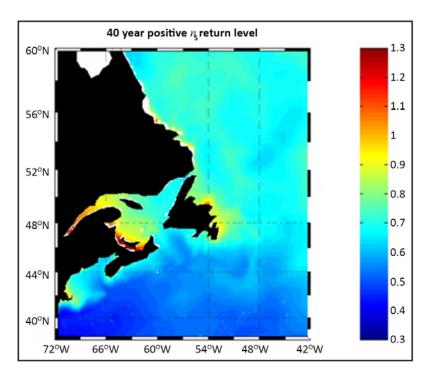


Figure 51. Storm surge maxima return level on the Atlantic coast of Canada based on the 40-year hindcast.

Hindcasting is a method of developing a model by testing it to see whether it accurately predicts past observations. The coloured bar indicates the 40-year positive surge return levels in metres. The most extreme surge events are expected to occur in the coastal regions highlighted by the warmest colours. Source: Bernier et al.,  $2006^{213}$ 

There were no comprehensive trend data on storm surges for the whole AME, however, a case study of storm surges in Charlottetown, PEI, indicate an increased severity and frequency in storm surge events between the 1940s and 1980s with surges over 90 cm becoming increasingly frequent (Figure 52). Since ice appears to have a damping effect on storm surge severity, storm surges and wave erosion may become more severe in a warmer climate, with reduced ice in the Gulf of St. Lawrence (see Ice across biomes section on page 37).<sup>107</sup>

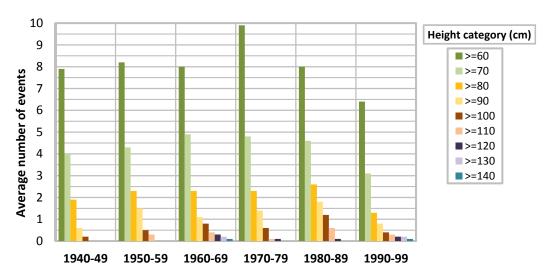


Figure 52. Average number of storm surges per year above the threshold, by decade, at Charlottetown, PEI, 1940s—1990s.

Source: adapted from Environment Canada in Forbes et al., 2006<sup>107</sup>

## Fire

Wildfires historically played an important role in forest dynamics in the AME, although at a much smaller scale than many other parts of Canada. <sup>214, 215</sup> Records from the 17th and 18th centuries suggest that lightning strikes regularly burned large areas of forest. <sup>216</sup> Today, fires are more numerous but smaller. The increased number of fires is due to the prevalence of human-caused fires. On average, between the 1960s and 2000s, 86% of fires were human-ignited. <sup>215</sup> However, the extent of area burned has been reduced through early detection and active fire suppression.

Since the 1950s, large forest fires (those greater than 2 km²) have not been a common or significant natural disturbance. From 1959 to 2007, an average of only 34 km² (0.02% of the AME) burned annually (Figure 53). Years with no large fires were common. Area burned was low due to fire prevention, early detection, and rapid suppression. Overall, total area burned was lower in the 1960s and 1970s, higher in the 1980s and 1990s, and lower again in the 2000s.<sup>215</sup>

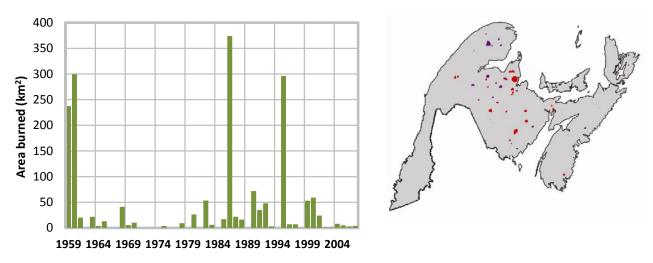


Figure 53. Total annual area burned by large fires (>2 km² in size) from 1959 to 2007 (left) and distribution of large fires from the 1980s to present in the Atlantic Maritime Ecozone $^+$ . On map, red is 1980s, purple is 1990s. Fires from 2000s (up to 2007 included) are too small to show. Source: Krezek-Hanes et al., 2010;<sup>215</sup> data from 1959–1994 from the large fire database (Stocks et al., 2003)<sup>217</sup> and data from 1995–2007 from remote sensing.

For some forests in the AME, fire has historically played an important role in the stand dynamics of forests, impacting tree species composition, age-class distribution, and patterns of succession.<sup>214, 218</sup> For example, in Nova Scotia, fire maintained forest diversity in pure Jack pine (*Pinus banksiana*) stands in Cumberland County and black spruce/white pine stands in the St. Mary's River area. Over time, fire suppression is expected to reduce forest diversity in these areas.<sup>219</sup>

In some ecosystems, repeated fire disturbance is important because it limits tree growth. Loss of soil fertility and hardpan formation in the soil profile caused by fires, combined with the allelopathic effect of heath-like vegetation on coniferous species, can create open woodland ecosystems with stunted trees, as in the barrens of southwest Nova Scotia. Natural fires have helped to maintain the Annapolis Valley heathlands.<sup>134</sup>

# Large-scale native insect outbreaks

Insect outbreaks are among the most frequent natural disturbances in the AME and the most common natural pathway for forest regeneration. Like fires, insect outbreaks also strongly influence a forest's successional dynamics (growth, in-growth, and mortality). However, unlike fires, insect outbreaks usually result in individual tree or small-patch replacement, rather than the loss of large stands. 16

#### Spruce budworm

Spruce budworm, which is native to North American boreal and mixedwood forests, is the most influential forest insect in the AME. 16, 221 Outbreaks occur somewhat synchronously over extensive areas, 222 but the duration of outbreaks varies regionally. Typically, periods of high

defoliation last 5 to 25 years.<sup>221, 223</sup> Recurring spruce budworm outbreaks play an important role in shaping forest ecosystems. They influence the residual forest stand's species composition, age-class distribution, successional dynamics, and forest condition.<sup>20, 223-225</sup> In addition, because spruce budworm and other insect outbreaks occur frequently and cover large areas, they affect the forest's carbon flux.<sup>224</sup>

There is no consensus on whether frequency and severity of outbreaks is changing. Some studies have found that the frequency of budworm outbreaks has increased,<sup>221, 222, 226</sup> while others have not found trends, especially when longer time scales were considered. For example, Boulanger and Arseneault<sup>223</sup> found that the outbreak frequency in eastern Quebec was stable from 1500 to 2000, with a return-interval of between 30 and 48 years (Figure 54).

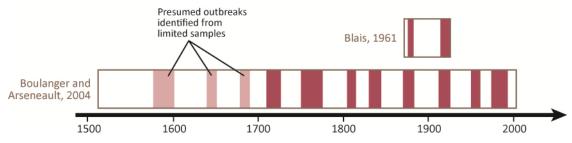
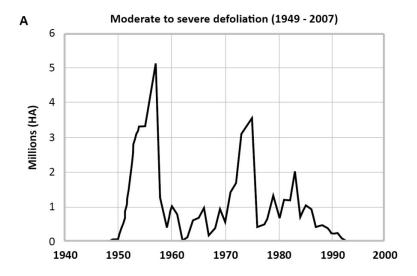


Figure 54. Intervals of spruce budworm outbreaks in eastern Quebec identified during previous reconstructions based on tree-ring chronologies, 1500–2000.

Source: adapted from Boulanger and Arseneault, 2004<sup>223</sup>

Some studies suggest that the severity of attacks increased during the 20<sup>th</sup> century.<sup>227, 228</sup> In contrast, severity of attacks decreased in New Brunswick from 1949 to 2007 (Figure 55).<sup>137</sup> This decline could have resulted from insecticide applications to combat spruce budworm outbreaks. Between 1972 and 1993, aerial insecticide was applied in New Brunswick on close to 50% of the moderately and severely infested areas, which reduced defoliation significantly.<sup>221</sup>



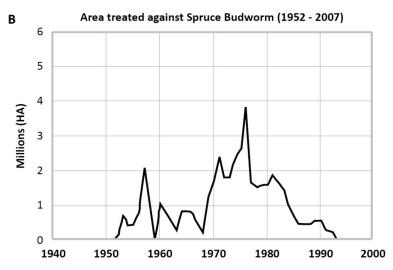


Figure 55. Trend in (A) spruce budworm defoliation in New Brunswick, 1949–2007, and (B) the area treated with pesticides, 1952–2007.

Source: modified from Carter et al., 2008<sup>137</sup>

Detectable changes to the severity and frequency of insect outbreaks across the range of the eastern spruce budworm were attributed to changes in forest harvest practices, reduced frequency of fire due largely to fire suppression, increased insecticide spraying, and less reliability in outbreak records reconstructed from historic periods. 16, 134, 223

Theme Habitat, wildlife, and ecosystem processes

# **Food webs**

## **National key finding**

Fundamental changes in relationships among species have been observed in marine, freshwater, and terrestrial environments. The loss or reduction of important components of food webs has greatly altered some ecosystems.

There was limited information on changes in trophic dynamics and population cycles in the AME. The loss or reduction of top mammalian predators can result in substantial changes to food webs. Historically, wolves were the largest mammalian predator but they were extirpated from New Brunswick and Nova Scotia sometime between 1870 and 1921.<sup>229</sup> American martens (*Martes americana*) were extirpated from southern Quebec and PEI, and black bear (*Ursus americanus*) and lynx (*Lynx canadensis*) were also extirpated from PEI.

At the same time, coyotes expanded their ranges into the AME (Figure 56) and replaced wolves as the top predator. Across eastern North America, including the AME, coyotes have exerted a strong "top-down effect" on forest ecosystems;<sup>230</sup> they directly reduced the abundance of prey which indirectly reduced the abundance of smaller carnivores such as red foxes (*Vulpes vulpes*).<sup>231</sup> By reducing the abundance of smaller carnivores, coyotes also indirectly increased the number of birds, creating a positive relationship between coyotes and scrub-bird populations.

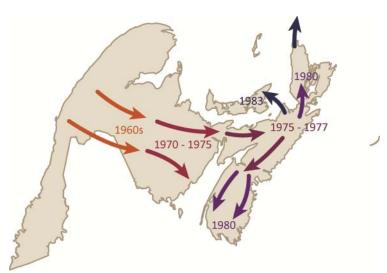


Figure 56. Chronology of colonization of a portion of the Atlantic Maritime Ecozone $^{+}$  by the eastern coyote from the 1960s to the 1980s.

Source: Moore and Parker, 1992<sup>232</sup>

# THEME: SCIENCE/POLICY INTERFACE

**Key finding 21** 

**Theme** Science/policy interface

# Biodiversity monitoring, research, information management, and reporting

#### National key finding

Long-term, standardized, spatially complete, and readily accessible monitoring information, complemented by ecosystem research, provides the most useful findings for policy-relevant assessments of status and trends. The lack of this type of information in many areas has hindered development of this assessment.

Although localized ecological studies and a few long-term data sets exist, information gaps made it difficult to determine ecological trends in the AME. Coordinated monitoring of biodiversity in the AME was generally lacking. More data were available for economically valuable species and biomes such as Atlantic salmon and forests, as well as certain species at risk such as woodland caribou. Information was lacking for protection and stewardship of private land.

# Strengths

- Population trends for migratory birds.
- Occurrence and population trends for Atlantic salmon.
- Some populations of invasive non-native species.

# Critical gaps identified

- Wetland trend data.
- Information on status and trends for non-vascular plants and invertebrates.
- Trend data on the coastal biome for the AME.
- Changes in trophic structures are poorly documented, particularly lacking for lakes.
- Ecological impacts of many invasive non-native species.

**Key finding 22** 

**Theme** Science/policy interface

# Rapid change and thresholds

#### National key finding

Growing understanding of rapid and unexpected changes, interactions, and thresholds, especially in relation to climate change, points to a need for policy that responds and adapts quickly to signals of environmental change in order to avert major and irreversible biodiversity losses.

Due to the underlying geology of the AME, aquatic ecosystems have less capacity to buffer acid and, consequently, a lower threshold for ecosystem damage from atmospheric acid deposition

than is found in other parts of Canada. As levels of acid deposition exceeded buffering capacity in lakes and rivers, Atlantic salmon populations and the number of salmon-bearing rivers in Nova Scotia rapidly declined. Moreover, impacts may persist longer than previously thought; rivers have not recovered with reductions in acid deposition, suggesting that these rivers are now in an alternative stable state (see Lakes and Rivers key finding on page 7).

Remaining forests in the AME are simpler, less diverse, and younger as a result of forest management practices (see Forests key finding on page 6). The landscape has also been highly fragmented by resource and road development, reducing the area of intact ecosystems. It is unclear to what extent these changes have contributed to the loss of native species, such as large mammals. However, in many parts of the AME, it is likely that a threshold of development and fragmentation has been reached where some species (e.g., caribou, black bear) could not survive if re-introduced.

One possible reason for rapid changes is that damage to ecosystems may accelerate because of the interaction of stressors. This is especially relevant for climate change. For example, coastal erosion in the AME is increasing, threatening wetlands, beach, and dune ecosystems. Development and hardening of the foreshore have made coastal ecosystems more susceptible to erosion. Rise in sea level, reduced sea ice, and more tropical storms in the Atlantic Ocean, all related to climate change (see Coastal key finding on page 7), accelerate the rate of erosion. Climate change may also make ecosystems more vulnerable to invasion from non-native species (e.g., warm-water fish species) and insect outbreaks, and increase the susceptibility of native species to diseases and infections.

# CONCLUSION: HUMAN WELL-BEING AND BIODIVERSITY

Over the last 400 years, the AME has had primarily a resource-based economy based on forestry, fishing, agriculture, mining, and, more recently, tourism. Although many of these industries are dependent on healthy, functioning ecosystems, industrial development has also had a large influence on the status and trends of AME ecosystems. In some cases, forestry, development, climate change, and acid deposition have impaired the ability of these ecosystems to continue to provide important goods and services.

Although forests still cover 80% of the landscape, forestry, fire suppression, and insect outbreaks have reduced species diversity, altered species composition, and shifted the age structure of forests towards younger stands. Remaining forests have also been highly fragmented by roads affecting forest-dwelling species. Caribou in the AME consist of a single, remnant endangered population and the moose population has declined. Many of the top mammalian predators, such as wolves, American marten, black bears, and lynx, were extirpated from all or most of the AME as a result of combined pressure from habitat changes and historic hunting.

Coastal ecosystems have also been impacted by industrial, urban, and cottage development. Some of the highest rates of wetland loss have been in coastal wetlands. The loss of beaches, dunes, and eelgrass meadows has reduced the suitability of coastlines as habitats for some species, such as shorebirds. Coastal ecosystem loss has also increased the vulnerability of the coastline to erosion from sea-level rise and storm surges, with associated hazards to human life and property. Port activities have introduced invasive non-native species, leading to declines particular tree species and economic impacts, for example, to timber production.

Freshwater lakes and rivers have been altered by climate-driven changes (e.g., changes in flow regimes, changes to ice freeze and thaw dates), the presence of dams, and excess nutrient runoff from agriculture. The AME has some of the most acid sensitive terrain in Canada and, as a result of historic acid deposition, many Atlantic salmon runs have been lost. Introduced fish species have altered food webs and aquatic community composition, as have didymo blooms. The impacts of climate change, although projected to be lower in the AME than other Canadian ecozones<sup>+</sup>, will exacerbate these changes.

Food production in the AME is mostly confined to PEI, Nova Scotia's Annapolis Valley, and New Brunswick's Saint John River Valley. Due to the expansion of cropland, agricultural land has become less capable of supporting wildlife. The AME has some of the highest residual soil nitrogen values and, due to its climate, some of highest soil erosion risks in Canada. However, soil erosion risk on agricultural lands in the AME has declined.

Changes to natural disturbance regimes include the suppression of fire, increased disturbance by extreme weather events, and insect outbreaks. As in many of the ecozones<sup>+</sup> in Canada, it is difficult to gauge the impacts to biodiversity, natural disturbances, and ecological processes due to a lack of comprehensive long-term monitoring.

# REFERENCES

- 1. Environment Canada. 2006. Biodiversity outcomes framework for Canada. Canadian Councils of Resource Ministers. Ottawa, ON. 8 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=F14D37B9-1">http://www.biodivcanada.ca/default.asp?lang=En&n=F14D37B9-1</a>.
- 2. Federal-Provincial-Territorial Biodiversity Working Group. 1995. Canadian biodiversity strategy: Canada's response to the Convention on Biological Diversity. Environment Canada, Biodiversity Convention Office. Hull, QC. 86 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=560ED58E-1">http://www.biodivcanada.ca/default.asp?lang=En&n=560ED58E-1</a>.
- 3. Federal, Provincial and Territorial Governments of Canada. 2010. Canadian biodiversity: ecosystem status and trends 2010. Canadian Councils of Resource Ministers. Ottawa, ON. vi + 142 p. http://www.biodivcanada.ca/default.asp?lang=En&n=83A35E06-1.
- 4. Eaton, S. 2013. Atlantic Maritime Ecozone<sup>+</sup> status and trends assessment. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Ecozone<sup>+</sup> Report. Canadian Councils of Resource Ministers. Ottawa, ON. Draft report.
- 5. Ecological Stratification Working Group. 1995. A national ecological framework for Canada. Agriculture and Agri-Food Canada, Research Branch, Centre for Land and Biological Resources Research and Environment Canada, State of the Environment Directorate, Ecozone Analysis Branch. Ottawa, ON/Hull, QC. vii + 125 p.
- 6. Rankin, R., Austin, M. and Rice, J. 2011. Ecological classification system for the ecosystem status and trends report. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 1. Canadian Councils of Resource Ministers. Ottawa, ON. ii + 14 p. http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-1.
- 7. Ahern, F., Frisk, J., Latifovic, R. and Pouliot, D. 2011. Monitoring ecosystems remotely: a selection of trends measured from satellite observations of Canada. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 17. Canadian Councils of Resource Ministers. Ottawa, ON. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 8. Latifovic, R. and Pouliot, D. 2005. Multitemporal land cover mapping for Canada: methodology and products. Canadian Journal of Remote Sensing 31:347-363.
- 9. Environment Canada. 2005. Ecozones of Canada. Environment Canada, State of the Environment Infobase. <a href="http://www.ec.gc.ca/soer-ree/English/vignettes/default.cfm">http://www.ec.gc.ca/soer-ree/English/vignettes/default.cfm</a>.
- 10. Canadian Council of Forest Ministers. 2006. Criteria and indicators of sustainable forest management in Canada: national status 2005. Canada Forest Service, Natural Resources Canada. Ottawa, ON. 154 p. + appendices.
- 11. Rothwell, N., Bollman, R.D., Tremblay, J. and Marshall, J. 2002. Recent migration patterns in rural and small town Canada. Agriculture and Rural Working Paper Series No. 55. Statistics Canada, Agricultural Division. Ottawa, ON. 71 p.
- 12. Environment Canada. 2009. Unpublished analysis of population data by ecozone<sup>+</sup> from: Statistics Canada Human Activity and the Environment Series, 1971-2006. Community

- profile data was used to make adjustments due to differences in the ecozone/ecozone+boundary.
- 13. Canadian Council of Forest Ministers. 2001. Canada's National Forest Inventory (CanFI) [online]. <a href="https://nfi.nfis.org/canfi.php?page=summaries&lang=en">https://nfi.nfis.org/canfi.php?page=summaries&lang=en</a>
- 14. Glen, W.M. 1997. Prince Edward Island 1935/1936 forest cover type mapping. Silviculture Development, Forestry Division, PEI Department of Agriculture and Forestry. Charlottetown, PEI. 9 p.
- 15. Mosseler, A., Lynds, J.A. and Major, J.E. 2003. Old-growth forests of the Acadian forest region. Environmental Reviews 11:S47-S77.
- 16. Loo, J. and Ives, N. 2003. The Acadian forest: historical condition and human impacts. The Forestry Chronicle 79:462-474.
- 17. Forest Stewardship Council Canada Working Group. 2007. Certification standards for best forestry practices in the Maritimes region. Forest Stewardship Council Canada. 91 p.
- 18. Rowe, J.S. 1972. Forest regions of Canada. Canadian Forestry Service Publication No. 1300. Publishing Division, Information Canada. Ottawa, ON. x + 172 p.
- 19. Lee, P., Gysbers, J.D. and Stanojevic, Z. 2006. Canada's forest landscape fragments: a first approximation (a Global Forest Watch Canada report). Global Forest Watch Canada. Edmonton, AB. 97 p.
- 20. Canadian Forest Service. 2012. Canada's forest regions. Canadian Forest Service, Natural Resources Canada. Map.
- 21. Natural Resources Canada. 2007. Forest ecosystems of Canada: forest regions classification [online]. Natural Resources Canada, Canadian Forest Service. <a href="http://ecosys.cfl.scf.rncan.gc.ca/classification/intro-region-eng.asp">http://ecosys.cfl.scf.rncan.gc.ca/classification/intro-region-eng.asp</a> (accessed 16 August, 2007).
- 22. Ministère des Ressources naturelles. 2005. Système hiérarchique de classification écologique du territoire. Direction des inventaires forestiers, Ministère des ressources naturelles, Gouvernement du Québec. Québec, QC.
- 23. Pannozza, L. and Colman, R. 2008. Genuine Progess Index (GPI) forest headline indicators for Nova Scotia. GPIAtlantic. Glen Haven, NS. vi + 54 p.
- 24. Ministère des Ressources naturelles et Faune. 2009. Statistiques forestières 2009. Updated from Ministère des Ressources naturelles, 2002. Unpublished data.
- 25. Ministère des Ressources naturelles. 2002. Rapport synthèse sur l'état des forêts Québécoises 1995-1999. Gouvernement du Québec. Charlesbourg, QC. 8 p.
- 26. Canadian Climate Impacts and Adaptation Research Network. 2007. Climate and deciduous forests in Maritime Canada [online]. Canadian Climate Impacts and Adaptation Research Network: Atlantic. <a href="http://www.c-ciarn.ca/pdf/factsheet\_deciduous\_forestry.pdf">http://www.c-ciarn.ca/pdf/factsheet\_deciduous\_forestry.pdf</a> (accessed 3 August, 2007).
- 27. Anderson, M.G., Vickery, B., Gorman, M., Gratton, L., Morrison, M., Maillet, J., Olivero, A., Ferree, C., Morse, D., Kehm, G., Rosalska, K., Khanna, S. and Bernstein, S. 2006. The Northern Appalachian/Acadian Ecoregion: conservation assessment status and trends:

- 2006. The Nature Conservancy Eastern Regional Science in collaboration with The Nature Conservancy of Canada: Atlantic and Quebec Regions. Boston, MA. 34 p.
- 28. Boucher, Y., Arseneault, D. and Sirois, L. 2009. La forêt préindustrielle du Bas-Saint-Laurent et sa transformation (1820-2000) : implications pour l'aménagement écosystémique. Le Naturaliste Canadien 133:60-69.
- 29. Jobin, B., Latendresse, C., Maisonneuve, C., Sebbane, A. and Grenier, M. 2007. Changements de l'occupation du sol dansle sud du Québec pour la période 1993-2001. Série de rapports techniques No. 483. Environnement Canada, Service canadien de la faune, région du Québec. Sainte-Foy, QC. 112 p.
- 30. Dettmers, R. 2004. Blueprint for the design and delivery of bird conservation in the Atlantic northern forest [online]. U.S. Fish and Wildlife Service. <a href="http://www.acjv.org/resources.htm">http://www.acjv.org/resources.htm</a> (accessed 15 December, 2009).
- 31. Busby, D., Austin-Smith, S., Curley, R., Diamond, A., Duffy, T., Elderkin, M., Makepeace, S., Diamond, D., Melanson, R., Staicer, C. and Whittam, B. 2006. Partners in Flight Canada: Maritime Canada landbird conservation plan. Technical Report Series No. 449. Environment Canada, Canadian Wildlife Service, Atlantic Region. Sackville, NB. 43 p.
- 32. COSEWIC. 2008. COSEWIC assessment and status report on the Canada warbler *Wilsonia canadensis* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. vi + 35 p.
- 33. Sleep, D.J.H., Drever, M.C. and Szuba, K.J. 2009. Potential role of spruce budworm in range-wide decline of Canada warbler. Journal of Wildlife Management 73:546-555.
- 34. Butcher, G.S. and Niven, D.K. 2007. Combining data from the Christmas Bird Count and the Breeding Bird Survey to determine the continental status and trends of North American birds. National Audubon Society. Ivyland, PA. 34 p.
- 35. Bird Studies Canada. 2008. Atlas of breeding birds of the Maritime provinces [online]. <a href="http://www.mba-aom.ca/english/index.html">http://www.mba-aom.ca/english/index.html</a> (accessed 23 October, 2009).
- 36. COSEWIC. 2009. COSEWIC assessment and status report on the Bicknell's thrush *Catharus bicknelli* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. vii + 44 p.
- 37. Rimmer, C.C., McFarland, K.P., Ellison, W.G. and Goetz, J.E. 2001. Bicknell's thrush (*Catharus bicknelli*). *In* The birds of North America online. Edited by Poole, A. Cornell Lab of Ornithology. Ithaca, NY. <a href="http://bna.birds.cornell.edu/bna/species/592">http://bna.birds.cornell.edu/bna/species/592</a>.
- 38. McFarland, K.P., Rimmer, C.C., Frey, S.J.K., Faccio, S.D. and Collins, B.B. 2008. Demography, ecology and conservation of Bicknell's thrush in Vermont, with a special focus on the northeastern highlands. Technical Report No. 08-03. Vermont Center for Ecostudies. Norwich, VT.
- 39. COSEWIC. 2009. COSEWIC assessment and status report on the Bicknell's thrush *Catharus bicknelli* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. 44 p.
- 40. Campbell, G., Whittam, B. and Robertson, G. 2008. High elevation landbird program 6-year report. Bird Studies Canada. Port Rowan, ON. 18 p.

- 41. King, D., Lambert, J., Buonaccorsi, J. and Prout, L. 2008. Avian population trends in the vulnerable montane forests of the Northern Appalachians, USA. Biodiversity and Conservation 17:2691-2700.
- 42. Downes, C., Blancher, P. and Collins, B. 2011. Landbird trends in Canada, 1968-2006. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 12. Canadian Councils of Resource Ministers. Ottawa, ON. x + 94 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 43. USGS Patuxent Wildlife Research Center. 2010. The North American Breeding Bird Survey [online]. U.S. Geological Survey, U.S. Department of the Interior. <a href="http://www.pwrc.usgs.gov/BBS/">http://www.pwrc.usgs.gov/BBS/</a>
- 44. Rosenberg, K.V. and Hodgman, T.P. 2000. Partners in Flight Landbird Conservation Plan: physiographic area 28: eastern spruce-hardwood forest. Cornell Lab of Ornithology. Ithaca, NY. 48 p.
- 45. Trombulak, S.C., Anderson, M.G., Baldwin, R.F., Beazley, K., Ray, J.C., Reining, C., Woolmer, G., Bettigole, C., Forbes, G. and Gratton, L. 2008. The Northern Appalachian/Acadian Ecoregion: priority locations for conservation action. Two Countries, One Forest Special Report No. 1. The Science Working Group of Two Countries, One Forest. Warner, NH. ii + 58 p.
- 46. Milton, R. 2008. Personal communication. Information on wetland mapping and assessment methodologies in Atlantic Canada. Nova Scotia Department of Natural Resources. Kentville, NS.
- 47. Nova Scotia Department of the Environment. 1998. The state of the Nova Scotia environment. Nova Scotia Department of the Environment. Halifax, NS. viii + 44 p.
- 48. Wiken, E., Smith, W.G.B., Cinq-Mars, J., Latsch, C. and Gauthier, D. 2003. Habitat integrity in Canada: wildlife conservation in Canada. Background paper for the national conference on guidelines and tools for the evaluation of Natura 2000 sites in France. Wildlife Habitat Canada. Montpellier, France, March 3-5, 2003. 30 p.
- 49. Rubec, C.D.A. and Hanson, A.R. 2009. Wetland mitigation and compensation: Canadian experience. Wetlands Ecology and Management 17:3-14.
- 50. Fast, M., Collins, B. and Gendron, M. 2011. Trends in breeding waterfowl in Canada. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 8. Canadian Councils of Resource Ministers. Ottawa, ON. v + 37 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 51. Lepage, C. and Bordage, D. 2003. The American black duck. Environment Canada, Canadian Wildlife Service, Quebec Region. Québec, QC.
- 52. Longcore, J.R., McAuley, D.G., Hepp, G.R. and Rhymer, J.M. 2000. American black duck (*Anas rubripes*). *In* The birds of North America online. Edited by Poole, A. Cornell Lab of Ornithology. Ithaca, NY. <a href="http://bna.birds.cornell.edu/bna/species/481">http://bna.birds.cornell.edu/bna/species/481</a>.
- 53. Petrie, M.J., Drobney, R.D. and Sears, D.T. 2000. Mallard and black duck breeding parameters in New Brunswick: a test of the reproductive rate hypothesis. Journal of Wildlife Management 64:832-838.

- 54. Mank, J.E., Carlson, J.E. and Brittingham, M.C. 2004. A century of hybridization: decreasing genetic distance between American black ducks and mallards. Conservation Genetics 5:395-403.
- 55. Kirby, R.E., Sargeant, G.A. and Shutler, D. 2004. Haldane's rule and American black duck x mallard hybridization. Canadian Journal of Zoology 82:1827-1831.
- 56. Rhymer, J.M. 2006. Extinction by hybridization and introgression in anatine ducks. Acta Zoologica Sinica 52:583-585.
- 57. Environment Canada. 1973. Inventory of Canadian freshwater lakes. Inland Waters Directorate, Water Resources Branch. Ottawa, ON. 34 p.
- 58. Burridge, M. and Mandrak, N. 2009. Ecoregion description: 118: Northeast US and Southeast Canada Atlantic drainages [online]. *In* Freshwater ecoregions of the world. The Nature Conservancy and the World Wildlife Fund. <a href="http://www.feow.org/ecoregion\_details.php?eco=118">http://www.feow.org/ecoregion\_details.php?eco=118</a> (accessed 21 February, 2009).
- 59. Burridge, M. and Mandrak, N. 2009. Ecoregion description: 115: Canadian Atlantic Islands [online]. *In* Freshwater ecoregions of the world. The Nature Conservancy and the World Wildlife Fund. <a href="http://www.feow.org/ecoregion\_details.php?eco=115">http://www.feow.org/ecoregion\_details.php?eco=115</a> (accessed 21 February, 2009).
- 60. Burridge, M. and Mandrak, N. 2009. Ecoregion description: 119: Scotia -- Fundy [online]. *In* Freshwater ecoregions of the world. The Nature Conservancy and the World Wildlife Fund. <a href="http://www.feow.org/ecoregion\_details.php?eco=119">http://www.feow.org/ecoregion\_details.php?eco=119</a> (accessed 21 February, 2009).
- 61. Environment Canada. 2010. Hydrology of Canada [online]. <a href="http://www.ec.gc.ca/rhc-wsc/default.asp?lang=En&n=E94719C8-1">http://www.ec.gc.ca/rhc-wsc/default.asp?lang=En&n=E94719C8-1</a> (accessed 1 March, 2009).
- 62. Cannon, A., Lai, T. and Whitfield, P. 2011. Climate-driven trends in Canadian streamflow, 1961-2003. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 19. Canadian Councils of Resource Ministers. Ottawa, ON. Draft report.
- 63. Monk, W.A. and Baird, D.J. 2011. Biodiversity in Canadian lakes and rivers. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 19. Canadian Councils of Resource Ministers. Ottawa, ON. Draft report.
- 64. Eaton, S.T. and Boates, J.S. 2002. Assessing the threats to coastal plain flora in the Tusket River watershed. Nova Scotia Department of Natural Resources. Kentville, NS.
- 65. Canadian Dam Association. 2003. Dams in Canada. International Commission on Large Dams (ICOLD). Montréal, QC. CD-ROM.
- 66. Hill, N.M. and Blaney, S. 2010. Exotic and invasive vascular plants of the Atlantic Maritime Ecozone. *In* Assessment of species diversity in the Atlantic Maritime Ecozone. Edited by McAlpine, D.F. and Smith, I.M. NRC Research Press. Ottawa, ON. Chapter 11. pp. 215-232.
- 67. COSEWIC. 2012. COSEWIC assessment and status report on the plymouth gentian *Sabatia kennedyana* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. xi + 46 p.

- 68. COSEWIC. 2012. COSEWIC assessment and status report on the pink coreopsis *coreopsis rosea* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. xii + 42 p.
- 69. Keddy, P.A. and Wisheu, I.C. 1989. Ecology, biogeography, and conservation of coastal plain plants: some general principles from the study of Nova Scotian wetlands. RHODORA 91:72-94.
- 70. Richardson, S.M., Hanson, J.M. and Locke, A. 2002. Effects of impoundment and water-level fluctuations on macrophyte and macroinvertebrate communities of a dammed tidal river. Aquatic Ecology 36:493-510.
- 71. Hanson, J.M. and Locke, A. 2000. The status of the dwarf wedgemussel, *Alasmidonta heterodon*, in Canada. Canadian Field-Naturalist 114:271-278.
- 72. Neily, P., McCurdy, D., Stewart, B. and Quigley, E. 2004. Coastal forest communities of the Nova Scotian eastern shore ecodistrict No. 74. Ecosystem Management Group, Forest Management Planning, Nova Scotia Department of Natural Resources. 26 p.
- 73. Daigle, R.J. (ed.). 2006. Impacts of sea-level rise and climate change on the coastal zone of southeastern New Brunswick. Environment Canada. Dartmouth, NS. 611 p.
- 74. Environment Canada. 2003. Salt marsh conservation and restoration [online]. Environment Canada. <a href="http://www.ns.ec.gc.ca/wildlife/salt\_marsh/toc\_e.html">http://www.ns.ec.gc.ca/wildlife/salt\_marsh/toc\_e.html</a> (accessed 7 March, 2008).
- 75. Hanson, A., Berube, D., Forbes, D., O'Carroll, S., Ollerhead, J. and Olsen, L. 2006. Impacts of sea-level rise and residential development on salt-marsh area in southeastern New Brunswick, 1944-2001. *In* Impacts of sea-level rise and climate change on the coastal zone of southeastern New Brunswick. Edited by Daigle, R.J. Environment Canada. Dartmouth, NS. Chapter 4.6.3. pp. 408-421.
- 76. CBCL Limited. 2009. The 2009 state of Nova Scotia's coast: technical report. Edited by Adamson, J., Conrad, T.S., Huston, J., Mitchell, D. and McKeane, S.W. Nova Scotia Department of Fisheries and Aquaculture and Nova Scotia Provincial Oceans Network. Halifax, NS. 245 p.
- 77. DFO. 2009. Does eelgrass (*Zostera marina*) meet the criteria as an ecologically significant species? Canadian Science Advisory Secretariat Science Advisory Report No. 2009/018. Department of Fisheries and Oceans. Moncton, NB. 11 p.
- 78. Waycott, M., Duarte, C.M., Carruthers, T.J.B., Orth, R.J., Dennison, W.C., Olyarnik, S., Calladine, A., Fourqurean, J.W., Heck, K.L., Hughes, A.R., Kendrick, G.A., Kenworthy, W.J., Short, F.T. and Williams, S.L. 2009. Accelerating loss of seagrasses across the globe threatens coastal ecosystems. Proceedings of the National Academy of Sciences of the United States of America 106:12377-12381.
- 79. Standing Committee on Fisheries and Oceans. 2008. Fifth report of the Standing Committee on Fisheries and Oceans to the House of Commons. Government of Canada. Ottawa, ON. 2 p. <a href="http://www2.parl.gc.ca/HousePublications/Publication.aspx?DocId=3562841&Language=E&Mode=1&Parl=39&Ses=2">http://www2.parl.gc.ca/HousePublications/Publication.aspx?DocId=3562841&Language=E&Mode=1&Parl=39&Ses=2</a>.

- 80. Hanson, A.R. 2004. Status and conservation of eelgrass (*Zostera marina*) in eastern Canada. Technical Report Series No. 412. Environment Canada, Canadian Wildlife Service, Atlantic Region. Sackville, NB. 40 p.
- 81. Short, F.T. 2008. Report to the Cree Nation of Chisasibi on the status of eelgrass in James Bay. Jackson Estuarine Laboratory. Durham, NH. 30 p.
- 82. Short, F.T. and Wyllie-Echeverria, S. 1996. Natural and human-induced disturbance of seagrasses. Environmental conservation 23:17-27.
- 83. Williams, S.L. 2007. Introduced species in seagrass ecosystems: status and concerns. Journal of Experimental Marine Biology and Ecology 350:89-110.
- 84. Sharp, G. and Semple, R. 2004. Status of eelgrass beds in south-western Nova Scotia. *In* Status and conservation of eelgrass (*Zostera marina*) in Eastern Canada. Edited by Hanson, A.R. Technical Report Series No. 412. Environment Canada, Canadian Wildlife Service, Atlantic Region. Sackville, NB. p. 8.
- 85. Seymour, N.R., Miller, A.G. and Garbary, D.J. 2002. Decline of Canada geese (*Branta canadensis*) and common goldeneye (*Bucephala clangula*) associated with a collapse of eelgrass (*Zostera marina*) in a Nova Scotia estuary. Helgoland Marine Research 56:198-202.
- 86. Chapman, A. and Smith, J. 2004. Quantifying the rapid decline of eelgrass beds on the eastern shore of Nova Scotia between 1992 and 2002. *In* Status and conservation of eelgrass (*Zostera marina*) in eastern Canada. Edited by Hanson, A.R. Technical Report Series No. 412. Environment Canada, Canadian Wildlife Service, Atlantic Region. Sackville, NB. p. 9.
- 87. Locke, A. and Hanson, J.M. 2004. Changes in eelgrass in southern Gulf of St. Lawrence estuaries. *In* Status and conservation of eelgrass (*Zostera marina*) in Eastern Canada. Edited by Hanson, A.R. Technical Report Series No. 412. Environment Canada, Canadian Wildlife Service, Atlantic Region. Sackville, NB. pp. 10-12.
- 88. Martel, M.-C., Provencher, L., Grant, C., Ellefsen, H.-F. and Pereira, S. 2009. Distribution and description of eelgrass beds in Quebec. Canadian Science Advisory Secretariat Research Document 2009/050. Department of Fisheries and Oceans. Ottawa, ON. viii + 37 p.
- 89. Morrison, R.I.G. 1977. Use of the Bay of Fundy by shorebirds. *In* Fundy tidal power and the environment: proceedings of a workshop on the environmental implications of Fundy tidal power. Wolfville, NS, 4-5 November, 1976. Edited by Daborn, G.R. Acadia University Institute. Wolfville, NS. pp. 187-199.
- 90. Morrison, R.I.G. and Harrington, B. 1979. Critical shorebird resources in James Bay and eastern North America. *In* Transactions of the 44th North American Wildlife Natural Resources Conference. Wildlife Management Institute. Washington, D.C. pp. 498-507.
- 91. Hicklin, P.W. 1987. The migration of shorebirds in the Bay of Fundy. Wilson Bulletin 99:540-570.
- 92. Morrison, R.I.G., Downes, C. and Collins, B. 1994. Population trends of shorebirds on fall migration in eastern Canada 1974-1991. Wilson Bulletin 106:431-447.

- 93. Morrison, R.I.G., Aubry, Y., Butler, R.W., Beyersbergen, G.W., Downes, C., Donaldson, G.M., Gratto-Trevor, C.L., Hicklin, P.W., Johnston, V.H. and Ross, R.K. 2001. Declines in North American shorebird populations. Wader Study Group Bulletin 94:34-38.
- 94. Morrison, R.I.G. and Hicklin, P. 2001. Recent trends in shorebird populations in the Atlantic Provinces. Bird Trends 8:16-19.
- 95. Hicklin, P.W. 2001. A comparison of roost counts in 1970s and 1990s in the Bay of Fundy. Bird Trends 8:19-21.
- 96. Bart, J., Brown, S., Harrington, B. and Morrison, R.I.G. 2007. Survey trends of North American shorebirds: population declines or shifting distribution? Journal of Avian Biology 38:73-82.
- 97. Gratto-Trevor, C., Morrison, R.I.G., Collins, B., Rausch, J. and Johnston, V. 2011. DO NOT USE USE 70039 Trends in Canadian shorebirds. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 13. Canadian Councils of Resource Ministers. Ottawa, ON. iv + 32 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 98. COSEWIC. 2007. COSEWIC assessment and status report on the red knot *Calidris canutus* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. viii + 58 p.
- 99. USFWS. 2009. Flyways [online]. US Fish and Wildlife Service. <a href="www.flyways.us">www.flyways.us</a> (accessed 23 March, 2009).
- 100. Eadie, J.M., Savard, J.P. and Mallory, M.L. 2000. Barrow's goldeneye (*Bucephala islandica*). *In* The birds of North America online. Edited by Poole, A. Cornell Lab of Ornithology. Ithaca, NY. <a href="http://bna.birds.cornell.edu/bna/species/548">http://bna.birds.cornell.edu/bna/species/548</a>.
- 101. Eastern Habitat Joint Venture. 2007. Five-year implementation plan. Eastern Habitat Joint Venture. Fredericton, NB. 66 p.
- 102. O'Carroll, S., Berube, D., Hanson, A., Forbes, D., Ollerhead, J. and Olsen, L. 2006. Temporal changes in beach and dune habitat in southeastern New Brunswick, 1944-2001. *In* Impacts of sea-level rise and climate change on the coastal zone of southeastern New Brunswick. Edited by Daigle, R.J. Environment Canada. Dartmouth, NS. Chapter 4.6.4. pp. 421-426.
- 103. Ferland, C.L. and Haig, S.M. 2002. 2001 International Piping Plover Census. U.S. Geological Survey, Forest and Rangeland Ecosystem Science Center. Corvallis, OR. 293 p.
- 104. Amirault-Langlais, D.L. 2012. Data on hatching success of piping plover in eastern Canada. Unpublished data.
- 105. Goossen, J.P. 2007. Hinterland who's who: piping plover (factsheet) [online]. Environment Canada and Canadian Wildlife Federation.

  <a href="http://www.hww.ca/assets/pdfs/factsheets/piper-plover-en.pdf">http://www.hww.ca/assets/pdfs/factsheets/piper-plover-en.pdf</a> (accessed 6 July, 2010).
- 106. Environment Canada. 2006. Recovery strategy for the piping plover (*Charadrius melodus melodus*) in Canada [Proposed]. Species at Risk Act Recovery Strategy Series. Environment Canada. Ottawa, ON. 40 p.

- 107. Forbes, D.L., Parkes, G.S. and Ketch, L.A. 2006. Sea-level rise and regional subsidence. *In* Impacts of sea level rise and climate change on the coastal zone of southeastern New Brunswick. Edited by Daigle, R.J. Environment Canada. Dartmouth, NS. Chapter 4.1. pp. 34-94.
- 108. Elliott-Smith, E., Haig, S.M. and Powers, B.M. 2009. Data from the 2006 International Piping Plover Census. U.S. Geological Survey Data Series 426. U.S. Geological Survey, U.S. Department of the Interior. Reston, VA. 332 p.
- 109. Government of New Brunswick. 2002. A coastal areas protection policy for New Brunswick. The Sustainable Planning Branch, New Brunswick Department of the Environment and Local Government. Fredericton, NB. 18 p.
- 110. Bourque, A. and Simonet, G. 2008. Quebec. *In* From impacts to adaptation: Canada in a changing climate 2007. Edited by Lemmen, D.S., Warren, F.J., Lacroix, J. and Bush, E. Government of Canada. Ottawa, ON. Chapter 5. pp. 171-226.
- 111. Shaw, J., Taylor, R.B., Forbes, D.L., Ruz, M.H. and Solomon, S. 1998. Sensitivity of the coasts of Canada to sea-level rise. Bulletin No. 505. Geological Survey of Canada. 79 p. + map.
- 112. McCulloch M.M., Forbes, D.L., Shaw, R.W. and CAFF-A041 Scientific Team. 2002. Coastal impact of climate change and sea-level rise on Prince Edward Island: synthesis report. Open File 4261. Geological Survey of Canada. 62 p. + CD-ROM.
- 113. Government of Prince Edward Island. 2003. State of the environment. PEI Department of Fisheries, Aquaculture and Environment. Charlottetown, PEI. ii + 61 p.
- 114. Marine Environmental Data Service. 2008. Canadian tides and water levels data archive [online]. Department of Fisheries and Oceans, Ottawa, ON. <a href="http://www.meds-sdmm.dfo-mpo.gc.ca/isdm-gdsi/twl-mne/index-eng.htm">http://www.meds-sdmm.dfo-mpo.gc.ca/isdm-gdsi/twl-mne/index-eng.htm</a>
- 115. Savard, J.-P., Bernatchez, P., Morneau, F., Saucier, F., Gachon, P., Senville, S., Fraser, C. and Jolivet, Y. 2008. Étude de la sensibilité des côtes et de la vulnérabilité des communautés du golfe du Saint-Laurent aux impacts des changements climatiques synthèse des résultats. Ouranos. Rimouski, QC. 48 p.
- 116. Bernatchez, P., Fraser, C., Friesinger, S., Jolivet, Y., Dugas, S., Drejza, S. and Morissette, A. 2008. Sensibilité des côtes et vulnérabilité des communautés du golfe du Saint-Laurent aux impacts des changements climatiques. Rapport de recherche remis au Consortium Ouranos et au FACC. Laboratoire de dynamique et de gestion intégrée des zones côtières, Université du Québec. Rimouski, QC. 256 p.
- 117. Forbes, D.L., Parkes, G.S., Manson, G.K. and Ketch, L.A. 2004. Storms and shoreline retreat in the southern Gulf of St. Lawrence. Marine Geology 210:169-204.
- 118. Bernatchez, P. and Dubois., J.M.M. 2008. Seasonal quantification of coastal processes and cliff erosion on fine sediments shoreline in a cold temperate climate, North Shore of the St. Lawrence Maritime Estuary, Quebec. Journal of Coastal Research 24:169-180.
- 119. Environment Canada. 2008. Changes in lake ice signal a changing climate. NatureWatch and Environment Canada. Ottawa, ON. 8 p.

- 120. Duguay, C.R., Prowse, T.D., Bonsal, B.R., Brown, R.D., Lacroix, M.P. and Ménard, P. 2006. Recent trends in Canadian lake ice cover. Hydrological Processes 20:781-801.
- 121. Magnuson, J.J., Robertson, D.M., Benson, B.J., Wynne, R.H., Livingstone, D.M., Arai, T., Assel, R.A., Barry, R.G., Card, V., Kuusisto, E., Granin, N.G., Prowse, T.D., Stewart, K.M. and Vuglinski, V.S. 2000. Historical trends in lake and river ice cover in the Northern Hemisphere. Science 289:1743-1746.
- 122. IceWatch. 2008. IceWatch Canada [online]. NatureWatch, Nature Canada. <a href="http://www.naturewatch.ca/english/icewatch/">http://www.naturewatch.ca/english/icewatch/</a> (accessed 1 December, 2008).
- 123. Williams, G.P. 1970. A note on the break-up of lakes and rivers as indicators of climate change. Atmosphere 8:23-24.
- 124. Prowse, T.D. and Culp, J.M. 2003. Ice breakup: a neglected factor in river ecology. Revue canadienne de génie civil 30:128-144.
- 125. Cunjak, R.A. and Newbury, R.W. 2005. Atlantic coast rivers of Canada. *In* Rivers of North America. Edited by Benke, A.C. and Cushing, C.E. Academic Press. San Diego, CA. pp. 939-980.
- 126. Senneville, S. and Saucier, F.J. 2007. Étude de sensibilité de la glace de mer au réchauffement climatique dans le golfe et l'estuaire du Saint-Laurent. Institut des Sciences de la Mer de Rimouski. Rimouski, QC. 30 p.
- 127. IUCN. 1994. Guidelines for protected area management categories. Commission on National Parks and Protected Areas with the assistance of the World Conservation Monitoring Centre, International Union for Conservation of Nature. Gland, Switzerland and Cambridge, UK. x + 261 p.
- 128. Environment Canada. 2009. Unpublished analysis of data by ecozone⁺ from: Conservation Areas Reporting and Tracking System (CARTS), v.2009.05 [online]. Canadian Council on Ecological Areas. <a href="http://ccea.org/en\_carts.html">http://ccea.org/en\_carts.html</a> (accessed 5 November, 2009).
- 129. CCEA. 2009. Conservation Areas Reporting and Tracking System (CARTS), v.2009.05 [online]. Canadian Council on Ecological Areas. <a href="http://ccea.org/en\_carts.html">http://ccea.org/en\_carts.html</a> (accessed 5 November, 2009).
- 130. Millennium Ecosystem Assessment. 2005. Ecosystems and human well-being: synthesis. A report of the Millennium Ecosystem Assessment. Island Press. Washington, DC. 137 p.
- 131. Blaney, S. 2001. Exotic and invasive plants in Maritime Canada. Blomidon Field-Naturalist Newsletter, Vol. 28, p. 6.
- 132. Blaney, S. 2008. Abundance of exotic species in the Maritime Provinces. Atlantic Canada Conservation Data Centre. Unpublished data.
- 133. Atlantic Canada Conservation Data Centre. 2011. Data on total number of native and non-native plant species in the Maritime Provinces, 2001. Unpublished data.
- 134. Neily, P.D., Quigley, E.J., Stewart, B.J. and Keys, K.S. 2007. Forest disturbance ecology in Nova Scotia. Nova Scotia Department of Natural Resources, Renewable Resources Branch, Forestry Division, Ecosystem Management Group. Truro, NS. 36 p. Draft report (February 7, 2007).

- 135. Forbes, G.J., McAlpine, D.F. and Scott, F.W. 2010. Mammals of the Atlantic Maritime Ecozone. *In* Assessment of species diversity in the Atlantic Maritime Ecozone. Edited by McAlpine, D.F. and Smith, I.M. NRC Research Press. Ottawa, ON. pp. 693-718.
- 136. Fundy Model Forest Network and INFOR Inc. 2007. Acadian forest management and restoration. Fundy Model Forest Network, Natural Resources Canada. 57 p.
- 137. Carter, N., Hartling, L., Lavigne, D., Gullison, J., O'Shea, D., Proude, J., Farquhar, R. and Winter, D. 2008. Preliminary summary of forest pest conditions in New Brunswick in 2007 and outlook for 2008. New Brunswick Department of Natural Resources, Forest Pest Management Section. Fredericton, NB. i + 15 p.
- 138. Canadian Food and Inspection Agency. 2005. 2005 plant protection survey report [online]. <a href="http://www.inspection.gc.ca/english/plaveg/pestrava/surv/sit2006e.shtml">http://www.inspection.gc.ca/english/plaveg/pestrava/surv/sit2006e.shtml</a> (accessed 1 February, 2009).
- 139. Nova Scotia Department of Natural Resources. 2008. State of the forest report 1995-2005: Nova Scotia forests in transition. Report FOR 2008-3. Nova Scotia Department of Natural Resources. 40 p.
- 140. Nova Scotia Department of Fisheries and Aquaculture. 2005. Species fact sheet: smallmouth bass [online]. **Nova Scotia Department of Fisheries and Aquaculture**. <a href="http://www.gov.ns.ca/fish/sportfishing/species/smb.shtml">http://www.gov.ns.ca/fish/sportfishing/species/smb.shtml</a> (accessed 1 February, 2009).
- 141. Government of New Brunswick. 2009. Smallmouth bass fact sheet [online]. Government of New Brunswick, Department of Natural Resources. <a href="http://www.gnb.ca/0078/publications/SmallmouthBass-e.pdf">http://www.gnb.ca/0078/publications/SmallmouthBass-e.pdf</a> (accessed 1 February, 2009).
- 142. Sabine, M.E.J. 2009. Smallmouth bass distribution in New Brunswick. Working paper presented at the DFO Regional Science Advisory Process. Moncton, NB, 27-28 January, 2009.
- 143. Nova Scotia Department of Agriculture and Fisheries. 2005. Nova Scotia trout management plan. Nova Scotia Department of Agriculture and Fisheries, Inland Fisheries Division. Halifax, NS. 45 p.
- 144. LeBlanc, J. 2009. Geographic distribution of smallmouth bass, *Micropterus dolomieui*, in Nova Scotia: history of early introductions and factors affecting current range. Canadian Science Advisory Secretariat Research Document 2010/028. Department of Fisheries and Oceans. Pictou, NS. iv + 25 p.
- 145. La ministre du Pêches et Océans Canada. 2013. Impacts potentiels de l'introduction d'achigan à petite bouche sur le saumon atlantique : Analyse des risques. DFO Canadian Science Advisory Secretariat Science Advisory Report No. 2009/003. Pêches et Océans Canada. Moncton, NB. 12 p.
- 146. MacRae, P.S.D. and Jackson, D.A. 2001. The influence of smallmouth bass (*Micropterus dolomieu*) predation and habitat complexity on the structure of littoral zone fish assemblages. Canadian Journal of Fisheries and Aquatic Sciences 58:342-351.
- 147. Environment Canada. 2008. Science and technology leaders: standing room only at Montreal workshop on didymo [online]. Environment Canada.

- http://www.ec.gc.ca/scitech/default.asp?lang=En&n=DD0250A7-1 (accessed 1 February, 2009).
- 148. MDDEP-MRNF. 2009. Didymo 2008 État de la situation. Quebec, Ministère du Développement durable, de l'Environnement et des Parcs, Ministère des Ressources naturelles et de la Faune. 5 p.
- 149. Gillis, C.A. and Chalifour, M. 2010. Changes in the macrobenthic community structure following the introduction of the invasive algae Didymosphenia geminata in the Matapedia River (Qu+¬bec, Canada). Hydrobiologia 647:63-70.
- 150. MDDEP-MRNF. 2007. What is didymo and how can we prevent it from spreading in our rivers? Québec, Ministère du Développement durable, de l'Environnement et des Parcs et Ministère des Ressources naturelles et de la Faune. 13 p.
- 151. Darling, J.A., Bagley, M.J., Roman, J., Tepolt, C.K. and Geller, J.B. 2008. Genetic patterns across multiple introductions of the globally invasive crab genus Carcinus. Molecular Ecology 17:4992-5007.
- 152. Blakeslee, A.M.H., McKenzie, C.H., Darling, J.A., Byers, J.E., Pringle, J.M. and Roman, J. 2010. A hitchhiker's guide to the Maritimes: anthropogenic transport facilitates long–distance dispersal of an invasive marine crab to Newfoundland. Diversity and Distributions 16:879-891.
- 153. DeGraaf, J.D. and Tyrrell, M.C. 2004. Comparison of the feeding rates of two introduced crab species, Carcinus maenas and Hemigrapsus sanguineus, on the blue mussel, Mytilus edulis. Northeastern Naturalist 11:163-167.
- 154. Klassen, G. and Locke, A. 2007. A biological synopsis of the European green crab, *Carcinus maenas*. Canadian Manuscript Report of Fisheries and Aquatic Sciences No. 2818. Fisheries and Oceans Canada, Gulf Fisheries Centre. Moncton, NB. vii + 75 p.
- 155. Davis, R.C., Short, F.T. and Burdick, D.M. 1998. Quantifying the effects of green crab damage to eelgrass transplants. Restoration Ecology 6:297-302.
- 156. Drury, C.F., Yang, J.Y. and De Jong, R. 2011. Trends in residual soil nitrogen for agricultural land in Canada, 1981-2006. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 15. Canadian Councils of Resource Ministers. Ottawa, ON. iii + 16 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 157. Commission on Nitrates in Groundwater. 2008. The report of the Commission on Nitrates in Groundwater. Government of Prince Edward Island. Charlottetown, PEI. 84 p.
- 158. PEI Department of Fisheries, Aquaculture and Environment. 2008. Data on nitrate levels in private wells on Prince Edward Island. Unpublished data.
- 159. Jiang, Y. 2009. Data on groundwater nitrate concentrations in watersheds on Prince Edward Island. Unpublished data.
- 160. Jiang, Y., Somers, G.H. and Mutch, J. 2004. Application of numerical modeling to groundwater assessment and management in Prince Edward Island. 57th Canadian Geotechnical Conference and the 5th Joint CGS/IAH-CNC Conference. Quebec City, QC, 24-27 October, 2004. Conference presentation.

- 161. Jiang, Y. and Somers, G. 2009. Modeling effects of nitrate from non-point sources on groundwater quality in an agricultural watershed in Prince Edward Island, Canada. Hydrogeology Journal 17:707-724.
- 162. Burkart, M.R. and James, D.E. 1999. Agricultural-nitrogren contributions to hypoxia in the Gulf of Mexico. Journal of Environmental Quality 28:850-859.
- 163. PEI Department of Environment, Fisheries and Forestry. 2009. Data on number of anoxic events on Prince Edward Island between 2002 and 2008. Unpublished data.
- 164. van Bochove, E., G.Thériault, Denault, J.-T., Dechmi, F., Rousseau, A.N. and Allaire, S.E. 2010. Phosphorus. *In* Environmental Sustainability of Canadian Agriculture: Agri-Environmental Indicator Report Series. Report No. 3. Edited by Eilers, W., MacKay, R., Graham, L. and Lefebvre, A. Agriculture and Agri-Food Canada. Ottawa, ON. pp. 87-93.
- 165. van Bochove, E., Thériault, G., Denault, J.T., Dechmi, F., Allaire, S.E. and Rousseau, A.N. 2012. Risk of phosphorus desorption from Canadian agricultural land: 25-year temporal trend. Journal of Environmental Quality 41:1402-1412.
- 166. Berryman, D. 2008. État de l'écosystème aquatique du bassin versant de la rivière Yamaska : faits saillants 2004-2006. Direction du suivi de l'état de l'environnement, ministère du Développement durable, de l'Environnement et des Parcs. Québec. 22 p.
- 167. Smith, V.H. 1983. Low nitrogen to phosphorus ratios favor dominance by blue-green algae in lake phytoplankton. Science 221:669-671.
- 168. Ministère du Développement durable, de l'Environnement et des Parcs. 2009. Bilan des lacs et cours d'eau touchés par une fleur d'eau d'algues bleu-vert au Québec. De 2004 à 2008 [online]. <a href="http://www.mddep.gouv.qc.ca/eau/algues-bv/bilan/liste-comparative.asp">http://www.mddep.gouv.qc.ca/eau/algues-bv/bilan/liste-comparative.asp</a> (accessed July, 2009).
- 169. Commission for Environmental Cooperation. 2008. The North American mosaic: an overview of key environmental issues. Communications Department of the CEC Secretariat. Montreal, QC. 62 p.
- 170. Clair, T.A., Dennis, I.F., Scruton, D.A. and Gilliss, M. 2007. Freshwater acidification research in Atlantic Canada: a review of results and predictions for the future. Environmental Reviews 15:153-167.
- 171. New England Governers/Eastern Canadian Premiers Forest Mapping Group. 2007. Acid rain action plan: action item 4: forest mapping project: mapping forest sensitivity to atmospheric acid deposition: 2006-2007 annual report. The Committee on the Environment of The Conference of New England Governors and Eastern Canadian Premiers. 12 p.
- 172. Duchesne, L., Ouimet, R. and Houle, D. 2002. Basal area growth of sugar maple in relation to acid deposition, stand health, and soil nutrients. Journal of Environmental Quality 31:1676-1683.
- 173. Driscoll, C.T., Lawrence, G.B., Bulger, A.J., Butler, T.J., Cronan, C.S., Eagar, C., Lambert, K.F., Likens, G.E., Stoddard, J.L. and Weathers, K.C. 2001. Acidic deposition in the northeastern United States: sources and inputs, ecosystem effects, and management strategies. BioScience 51:180-198.

- 174. Nova Scotia Salmon Federation. 2009. Nova Scotia salmon federation news and issues: acid rain [online]. Nova Scotia Salmon Federation.

  <a href="http://www.novascotiasalmon.ns.ca/newsandissues/acidrain.htm">http://www.novascotiasalmon.ns.ca/newsandissues/acidrain.htm</a> (accessed 1 March, 2009).
- 175. Watt, W.D., Scott, C.D., Zamora, P.J. and White, W.J. 2000. Acid toxicity levels in Nova Scotian rivers have not declined in synchrony with the decline in sulfate levels. Water, Air, and Soil Pollution 118:203-229.
- 176. Clair, T.A., Dennis, I.F., Amiro, P.G. and Cosby, B.J. 2004. Past and future chemistry changes in acidified Nova Scotian Atlantic salmon (*Salmo salar*) rivers: a dynamic modeling approach. Canadian Journal of Fisheries and Aquatic Sciences 61:1965-1975.
- 177. Dennis, I.F., Clair, T.A. and Kidd, K. 2012. The distribution of dissolved aluminum in Atlantic salmon (*Salmo salar*) rivers of Atlantic Canada and its potential effect on aquatic populations. Canadian Journal of Fisheries and Aquatic Sciences 69:1174-1183.
- 178. Zhang, X., Brown, R., Vincent, L., Skinner, W., Feng, Y. and Mekis, E. 2011. Canadian climate trends, 1950-2007. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 5. Canadian Councils of Resource Ministers. Ottawa, ON. iv + 21 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 179. Moron, V., Vautard, R. and Ghil, M. 1998. Trends, inderdecadal and interannual oscillations in global sea surface temperatures. Climate Dynamics 14:545-569.
- 180. Barrow, E., Maxwell, B. and Gachon, P. (eds.). 2004. Climate variability and change in Canada: past, present and future. Environment Canada. Toronto, ON. 114 p.
- 181. Lemmen, D.S. and Warren, F.J. (eds). 2004. Climate change impacts and adaptation: a Canadian perspective. Government of Canada. Ottawa, ON. 174 p.
- 182. McKenzie, K. and Parlee, K. 2003. The road ahead adapting to climate change in Atlantic Canada [online]. Canadian Climate Impacts and Adaptation Research Network (C-CAIRN). <a href="http://www.elements.nb.ca/theme/climate03/cciarn/adapting.htm">http://www.elements.nb.ca/theme/climate03/cciarn/adapting.htm</a> (accessed 12 January, 2008).
- 183. Javorek, S.K., Antonowitsch, R., Callaghan, C., Grant, M. and Wiens, T. 2005. Wildlife habitat on farmland. *In* Environmental Sustainability of Canadian Agriculture: Agri-Environmental Indicator Report Series. Report No. 2. Edited by Lefebvre, A., Eilers, W. and Chunn, B. Agriculture and Agri-Food Canada. Ottawa, ON. pp. 158-164.
- 184. Javorek, S.K. and Grant, M.C. 2011. Trends in wildlife habitat capacity on agricultural land in Canada, 1986-2006. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 14. Canadian Councils of Resource Ministers. Ottawa, ON. vi + 46 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 185. McConkey, B.G., Lobb, D.A., Li, S., Black, J.M.W. and Krug, P.M. 2011. Soil erosion on cropland: introduction and trends for Canada. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 16. Canadian Councils of Resource Ministers. Ottawa, ON. iv + 22 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.

- 186. COSEWIC. 2002. COSEWIC assessment and update status report on the woodland caribou *Rangifer tarandus caribou* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. xi + 98 p.
- 187. Nova Scotia Department of Natural Resources. 2009. Nova Scotia endangered species act: legally listed species at risk [online]. Biodiversity Program, Nova Scotia Department of Natural Resources. <a href="http://www.gov.ns.ca/natr/wildlife/biodiversity/species-list.asp">http://www.gov.ns.ca/natr/wildlife/biodiversity/species-list.asp</a>
- 188. Dauphine, T.C. 1975. The disappearance of caribou reintroduced to Cape Breton Highlands National Park. Canadian Field-Naturalist 89:299-310.
- 189. Gaspésie Woodland Caribou Recovery Team. 2007. Gaspésie woodland caribou (*Rangifer tarandus caribou*) recovery plan (2002-2012): update. Ministère des Ressources naturelles et de la Faune. Secteur Faune Québec. xi + 57 p.
- 190. Nova Scotia Department of Natural Resources. 2003. Species at risk in Nova Scotia [online]. Government of Nova Scotia. <a href="http://www.gov.ns.ca/natr/wildlife/endngrd/#2003">http://www.gov.ns.ca/natr/wildlife/endngrd/#2003</a> (accessed 17 March, 2008).
- 191. Miramichi River Environmental Assessment Committee and Atlantic Coastal Action Program. 2007. State of the environment report for the Miramichi watershed [online]. Miramichi River Environmental Assessment Committee and the Atlantic Coastal Action Program. <a href="http://www.mreac.org/Publications/newsletterissue12.pdf">http://www.mreac.org/Publications/newsletterissue12.pdf</a> (accessed 6 July, 2007).
- 192. Gibson, J., Hubley, B., Chaput, G., Dempson, J.B., Caron, F. and Amiro, P. 2006. Summary of status and abundance trends for eastern Canadian Atlantic salmon (*Salmo salar*) populations. Canadian Science Advisory Secretariat Research Document 2006/026. Department of Fisheries and Oceans. 31 p.
- 193. COSEWIC. 2010. COSEWIC assessment and status report on the Atlantic Salmon *Salmo salar* (Nunavik population, Labrador population, Northeast Newfoundland population, South Newfoundland population, Southwest Newfoundland population, Northwest Newfoundland population, Quebec Eastern North Shore population, Quebec Western North Shore population, Anticosti Island population, Inner St. Lawrence population, Lake Ontario population, Gaspé-Southern Gulf of St. Lawrence population, Eastern Cape Breton population, Nova Scotia Southern Upland population, Inner Bay of Fundy population, Outer Bay of Fundy population) in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. xvii + 136 p.
- 194. Guignion, D. 2009. A conservation strategy for Atlantic salmon in Prince Edward Island. Prince Edward Island Council of the Atlantic Salmon Federation. Prince Edward Island. xiii + 168 p.
- 195. Ontario Ministry of Natural Resources. 2010. State of resources reporting, American eel in Ontario, February 2007 report and 2010 update. Government of Ontario. Peterborough, ON. 6 p.
- 196. COSEWIC. 2012. COSEWIC assessment and status report on the American eel *Anguilla rostrata* in Canada. Committee on the Status of Endangered Wildlife in Canada. Ottawa, ON. xii + 109 p.

- 197. de Lafontaine, Y., Gingras, F., Labonte, D., Marchand, F., La Croix, E. and Lagace, M. 2009. Decline of the American eel in the St. Lawrence River: effects of local hydroclimatic conditions on CPUE Indices. American Fisheries Society Symposium No. 58. Edited by Casselman, J. and Cairns, D.K. 228 p.
- 198. Jelks, H.L., Walsh, J., Burkhead, N.M., Contreras-Balderas, S., Díaz-Pardo, E., Hendrickson, D.A., Lyons, J., Mandrak, N.E., McCormick, F., Nelson, J.S., Platania, S.P., Porter, B.A., Renaud, C.B., Schmitter-Soto, J.J., Taylor, E.B. and Warren, Jr.M.L. 2008. Conservation status of imperiled North American freshwater and diadromous fishes. Fisheries 33:372-407.
- 199. Pouliot, D., Latifovic, R. and Olthof, I. 2009. Trends in vegetation NDVI from 1 km Advanced Very High Resolution Radiometer (AVHRR) data over Canada for the period 1985-2006. International Journal of Remote Sensing 30:149-168.
- 200. Gilabert, M.A., Gandia, S. and Melia, J. 1996. Analyses of spectral biophysical relationships for a corn canopy. Remote Sensing of Environment 55:11-20.
- 201. Santin-Janin, H., Garel, M., Chapuis, J.-L. and Pontier, D. 2009. Assessing the performance of NDVI as a proxy for plant biomass using non-linear models: a case study on the Kerguelen Archipelago. Polar Biology 32:861-871.
- 202. Myneni, R.B., Keeling, C.D., Tucker, C.J., Asrar, G. and Nemani, R.R. 1997. Increased plant growth in the northern high latitudes from 1981 to 1991. Nature 386:698-702.
- 203. Slayback, D.A., Pinzon, J.E., Los, S.O. and Tucker, C.J. 2003. Northern Hemisphere photosynthetic trends 1982-99. Global Change Biology 9:1-15.
- 204. Zhou, L., Tucker, C.J., Kaufmann, R.K., Slayback, D., Shabanov, N.V. and Myneni, R.B. 2001. Variations in northern vegetation activity inferred from satellite data of vegetation index during 1981 to 1999. Journal of Geophysical Research-Atmospheres 106:20,069-20,083.
- 205. Tateishi, R. and Ebata, M. 2004. Analysis of phenological change patterns using 1982-2000 Advanced Very High Resolution Radiometer (AVHRR) data. International Journal of Remote Sensing 25:2287-2300.
- 206. Goetz, S., Bunn, A.G., Fiske, G.J., Houghton, R.A. and Woodwell, G.M. 2005. Satellite-observed photosynthetic trends across boreal North America associated with climate and fire disturbance. Proceedings of the National Academy of Sciences of the United States of America 102:13521-13525.
- 207. Bunn, A.G. and Goetz, S.J. 2006. Trends in satellite-observed circumpolar photosynthetic activity from 1982 to 2003: the influence of seasonality, cover type, and vegetation density. Earth Interactions 10:1-19.
- 208. Vasseur, L. and Catto, N. 2008. Atlantic Canada. *In* From impacts to adaptation: Canada in a changing climate 2007. Edited by Lemmen, D.S., Warren, F.J., Lacroix, J. and Bush, E. Government of Canada. Ottawa, ON. Chapter 4. pp. 119-170.
- 209. Elsner, J.B., Kossin, J.P. and Jagger, T.H. 2008. The increasing intensity of the strongest tropical cyclones. Nature 455:92-95.

- 210. Saunders, M.A. and Lea, A.S. 2008. Large contribution of sea surface warming to recent increase in Atlantic hurricane activity. Nature 451:557-561.
- 211. Environment Canada. 2002. Atlantic hurricanes breaking records. Science and the Environment Bulletin 31:4-5.
- 212. Emanuel, K. 2005. Increasing destructiveness of tropical cyclones over the past 30 years. Nature 436:686-688.
- 213. Bernier, N., MacDonald, J., Ou, J., Ritchie, H. and Thompson, K. 2006. Storm-surge and meterological modelling. *In* Impacts of sea-level rise and climate change on the coastal zone of southeastern New Brunswick. Environment Canada. Ottawa, ON. Chapter 4. pp. 263-298.
- 214. Meades, B., Stewart, C., Goltz, J., MacQuarrie, K., Kavanagh, K., Sims, M. and Mann, G. 1999. Appendix F: ecoregion descriptions: Gulf of St. Lawrence lowland forests. *In* Terrestrial ecoregions of North America: a conservation assessment. Edited by Ricketts, T.H., Dinerstein, E., Olson, D.M. and Loucks, C.J. Island Press. Washington, DC. pp. 175-178.
- 215. Krezek-Hanes, C.C., Ahern, F., Cantin, A. and Flannigan, M.D. 2011. Trends in large fires in Canada, 1959-2007. Canadian Biodiversity: Ecosystem Status and Trends 2010, Technical Thematic Report No. 6. Canadian Councils of Resource Ministers. Ottawa, ON. v + 48 p. <a href="http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0">http://www.biodivcanada.ca/default.asp?lang=En&n=137E1147-0</a>.
- 216. Braune, B. 2008. NCP synopsis. Technical Background Reports: Ecosystem Status and Trends Report for Canada (in preparation).
- 217. Stocks, B.J., Mason, J.A., Todd, J.B., Bosch, E.M., Wotton, B.M., Amiro, B.D., Flannigan, M.D., Hirsch, K.G., Logan, K.A., Martell, D.L. and Skinner, W.R. 2003. Large forest fires in Canada, 1959-1997. Journal of Geophysical Research 108:8149-8161.
- 218. Forbes, G., Veen, H., Loo, J., Zelazny, V. and Woodley, S. 1998. Ecological change in the Greater Fundy ecosystem. *In* State of the Greater Fundy Ecosystem. Edited by Woodley, S., Forbes, G. and Skibicki, A. University of New Brunswick, Faculty of Forestry and Environmental Management. Fredericton, NB. Chapter 3. pp. 55-74.
- 219. Basquill, S.P., Woodley, S.J. and Purdy, A.B. 2001. The history and ecology of fire in Kejimkujik National Park. Technical Reports in Ecosystem Science No. 29. Parks Canada. Ottawa, ON. 189 p.
- 220. Colford-Gilks, A.K., D.A.MacLean, J.A.Kershaw, J. and M.Béland. 2012. Growth and mortality of balsam fir- and spruce-tolerant hardwood stands as influenced by stand characteristics and spruce budworm defoliation. Forest Ecology and Management 280:82-92.
- 221. Gray, D.R. and MacKinnon, W.E. 2007. Historical spruce budworm defoliation records adjusted for insecticide protection in New Brunswick, 1965-1992. Journal of the Acadian Entomological Society 3:1-6.
- 222. Royama, T., MacKinnon, W.E., Kettela, E.G., Carter, N.E. and Hartling, L.K. 2005. Analysis of spruce budworm outbreak cycles in New Brunswick, Canada, since 1952. Ecology 86:1212-1224.

- 223. Boulanger, Y. and Arseneault, D. 2004. Spruce budworm outbreaks in eastern Quebec over the last 450 years. Canadian Journal of Forest Research 34:1035-1043.
- 224. Gray, D.R. and MacKinnon, W.E. 2006. Outbreak patterns of the spruce budworm and their impacts in Canada. The Forestry Chronicle 82:550-561.
- 225. Colford, A. 2007. Effects of spruce budworm outbreaks on stand dynamics of balsam fir and red spruce tolerant hardwood mixedwoods. Thesis (M.Sc. Proposal Presentation). University of New Brunswick, Faculty of Forestry and Environmental Management. Fredericton, NB. 8 p.
- 226. Blais, J.R. 1983. Trends in the frequency, extent and severity of spruce budworm outbreaks in eastern Canada. Canadian Journal of Forest Research 13:539-545.
- 227. MacLean, D.A. 1994. Effects of spruce budworm outbreaks on the productivity and stability of balsam fir forests. The Forestry Chronicle 60:273-279.
- 228. Kettela, E.G. 1983. A cartographic history of spruce budworm defoliation from 1967 to 1981 in eastern North America. Information Report No. DPC-X-14. Environment Canada, Canadian Forestry Service. Hull, QC. 9 p.
- 229. Paquet, P.C. and Carbyn, L.N. 2003. Gray wolf. *In* Wild mammals of North America: biology, management and conservation. Edited by Feldhamer, G.A., Thompson, B.C. and Chapman, J.A. John Hopkins University Press. Baltimore, MD. Chapter 23. 1216.
- 230. Kays, R.W., Gompper, M.E. and Ray, J.C. 2008. Landscape ecology of eastern coyotes based on large-scale estimates of abundance. Ecological Applications 18:1014-1027.
- 231. Gompper, M.E. 2002. The ecology of northeast coyotes: current knowledge and priorities for future research. WCS Working Paper No. 17. Wildlife Conservation Society. Bronx, NY. 49 p.
- 232. Moore, G.C. and Parker, G.R. 1992. Colonization by the eastern coyote (*Canis latrans*). *In* Ecology and management of the eastern coyote. Edited by Boer, A.H. Wildlife Research Unit, University of New Brunswick. Fredericton, NB. pp. 23-37.